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1 Whale baleen as a biomonitor for per- and polyfluoroalkyl substances (PFAS)

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22 Abstract

23 Per- and polyfluoroalkyl substances (PFAS) comprise > 10,000 synthetic compounds that are
24 globally distributed and highly persistent but remain challenging to monitor. Here we assess the
25 utility of baleen—an accreting, keratinaceous tissue that baleen whales use for filter-feeding—to
26 track PFAS dynamics in marine food webs. In six species investigated, PFAS were detected in all
27 baleen tested (n = 18 plates, 220 samples, Σ_{10} PFAS range 0.02 - 60.5 ng/g dry weight), higher than
28 other tissue types besides liver. Three of the species in our dataset had not been tested for PFAS-
29 contamination previously and two of those species—blue whale and North Atlantic right whale—
30 are endangered species internationally. Apparent links were observed between PFAS and life-
31 history events by testing successive subsamples along the growth axis of the baleen plates. These
32 results establish baleen as a viable sample matrix for assessing PFAS contamination in marine
33 ecosystems by enabling multiyear time-series analyses through single-tissue sampling with
34 seasonal resolution.

36 Keywords

37 PFAS, PFOS, bioaccumulation, ecosystem sentinels, marine mammals, baleen whales

39 Synopsis

40 Baleen exhibits high affinity for PFAS, with concentrations varying along the growth axis,
41 enabling time-series PFAS monitoring in baleen whales.

42 **Introduction**

43 Synthetic chemical pollution is a potential driver of planetary change^{1,2}, necessitating
44 monitoring and mitigation to identify and minimize consequences for species and ecosystems. The
45 rapid introduction of novel substances, in concert with multiple-anthropogenic stressors (e.g.,
46 habitat degradation, climate change), can cause profound transformations to the biosphere³⁻⁵.
47 PFAS have been detected in fish and other aquatic wildlife for decades⁶, and some PFAS
48 bioaccumulate and can biomagnify within aquatic trophic pyramids⁷⁻⁹. Among marine vertebrates,
49 PFAS have been detected in the feathers of sea eagles¹⁰, and in the liver, muscle, and adipose
50 tissues of fish¹¹, seabirds^{12,13}, and marine mammals^{14,15}. Cetaceans are relatively well-studied in
51 regards to PFAS pollution¹⁵⁻²⁰; however, this research is limited by accessibility and viability of
52 blood-fed or protein-rich tissues to monitor PFAS over time.

53 Baleen whales (Mysticeti), the largest cetaceans, are sentinels of global change and marine
54 pollution²¹⁻²³. They feed on schooling forage fish or swarming pelagic crustaceans via bulk
55 filtration feeding. In this process, large volumes of prey-laden water are drawn into the buccal
56 cavity and filtered out using baleen plates attached to the upper palate²⁴. Baleen is an epidermal
57 tissue made of two morphologically distinct layers of alpha-keratin²⁵. As the baleen grows from
58 the gum tissue (i.e., the *Zwischensubstanz*)²⁶, it continually erodes on the lingual side, creating a
59 dense fringe mat, which acts as the filter. As the plates grow dorsoventrally, at a rate of ~20 cm
60 per year (range: 12-32 cm per year)²⁵, the chemical signature acquired when the baleen first
61 emerged is preserved, as in human hair²⁷⁻²⁹. This makes baleen a useful tissue for investigating
62 key physiological and phenological variables over time, including individual movements and diet
63 patterns via stable isotopes³⁰⁻³², hormonal changes related to breeding and feeding³³⁻³⁶, and
64 variation in exposure to trace metal pollution³⁷. However, only a limited time window for analysis

65 is available (the most recent 3-13 years)²⁵ as the plate eventually erodes distally and is replaced by
66 new keratinous tissue.

67 Despite its utility as a matrix for trace metal analysis^{37,38}, baleen has been underutilized for
68 organic pollutant monitoring, primarily due to concerns regarding its affinity for lipophilic
69 persistent organic pollutants (POPs). However, PFAS readily associate with specific proteins in
70 biological matrices, with demonstrated affinity for human liver fatty acid binding protein and
71 serum albumin, among others^{39,40}. PFAS have also been found in human hair, which is primarily
72 composed of keratin⁴¹. Here, we test the hypotheses that PFAS 1) are detectable in mysticete
73 baleen, 2) vary within an individual along a baleen plate, and 3) whether PFAS varies between
74 tissue types (baleen as compared to skin, blubber, liver, and gum tissues). We test these hypotheses
75 on samples from six species (Table 1). Unlike wet tissues, baleen can be stored at room temperature
76 without chemicals or fixatives. Thus, if PFAS are consistently detected in baleen, this may
77 facilitate pollutant monitoring at spatial and temporal scales previously unexplored by tapping into
78 the vast stores of baleen currently held by museums, stranding networks, and private or
79 opportunistic collections.

80

81 **Materials and Methods**

82 *Sample collection and storage*

83 U.S. West Coast samples were collected by The Marine Mammal Center (TMMC) under
84 National Oceanic and Atmospheric Administration (NOAA)/National Marine Fisheries Services
85 (NMFS) permit 18786-04 and 24359 (Fig. S1). In the Southern Gulf of Maine, samples were
86 collected by the International Fund for Animal Welfare (IFAW) under a NOAA/NMFS Permit No.
87 18786-06 (Fig. S1). Samples were collected in the Northern Maine region by the College of the

88 Atlantic's (COA) Marine Mammal Stranding Response Program (MMSRP) known as Allied
89 Whale under a stranding agreement or Permit No. 22723 NOAA/NMFS (Fig. S1). MMSRP
90 contributed one plate initially collected by Marine Animal Lifeline, a stranding network no longer
91 in existence, and two plates received from IFAW that were archived at COA. All samples were
92 collected during necropsy examinations from deceased whales following existing standard
93 necropsy procedures^{42,43}. Carcasses were either fresh dead or in moderate states of decomposition
94 at the time of sample collection (Table S1). Liver, blubber, and skin samples were stored in sterile
95 polypropylene bags at either -20°C or -80°C. Baleen was stored at room temperature or -20°C or
96 -80°C if fresh gum tissue was attached. Species included in this study were humpback whale
97 (*Megaptera novaeangliae*), common minke whale (*Balaenoptera acutorostrata*), North Atlantic
98 right whale (*Eubalaena glacialis*), sei whale (*B. borealis*), fin whale (*B. physalus*), and blue whale
99 (*B. musculus*) (Table S1).

100

101 *Sampling of baleen plates*

102 The exterior of each baleen plate was rinsed with PFAS-free milliQ water three times and
103 scrubbed as needed to remove dirt and dust. A Kimwipe was then soaked in acetone and used to
104 thoroughly wipe the plate to desorb any superficial organic contaminants; multiple Kimwipes were
105 used dependent on the size of the plate. The plate exterior was then triple-rinsed with LCMS-grade
106 methanol via a squirt bottle and allowed to dry in a fume hood before further manipulation. Baleen
107 plates were and subsampled sequentially and vertically every ~2 cm, beginning at the dorsal buccal
108 edge (Fig. 1), using an electric rotary grinder (Dremel model 4000) with a 1.25 cm sanding band
109 attachment to grind baleen into a fine powder. Sampling equipment was also wiped with Kimwipes
110 soaked in > 99% methanol to prevent cross-contamination. Overall, 221 discrete baleen samples

111 (~0.25 g powder per sample) were collected from 18 baleen plates from 17 different individuals
112 (Tables S1-S2, Supporting Information). Given the restrictions on collection and use of marine
113 mammal parts based on international and national law, this approach afforded a large sample size
114 unusual in contaminant studies relying on opportunistic samples from large whales.

115

116 *PFAS Analysis*

117 Liver and other wet tissue samples were spiked with internal standard and solvent-extracted
118 in acetonitrile using sonication, centrifugation, and freezing, paired with graphitized non-porous
119 carbon clean-up via a modified version of the extraction described in Spaan et al.¹⁵. Dry baleen
120 samples were spiked with 5ng of a mixture of mass-labeled surrogates and extracted twice in series
121 using a basic methanol digestion, sonication, centrifugation, and graphitized carbon clean-up
122 following a modified version of an extraction developed for feathers¹⁰ (for details, see Supporting
123 Information). Measurement and quantification of target PFAS in this proof-of-concept application
124 was achieved using liquid chromatography tandem-mass spectrometry (UPLC-MS/MS)
125 experiments in negative electrospray ionization mode (Supporting Information for details; Tables
126 S3-S9).

127

128 *Quality assurance and statistical analysis*

129 Procedural blanks were included with the sample set to monitor process recovery and
130 background contamination (Tables S10-S13). Baleen samples displayed an average median
131 recovery of 86% and wet tissue samples displayed an average median recovery of 76% (excluding
132 non-detected FTS, PFBA, PFBS). Method detection limits (MDLs) were defined as procedural
133 blank levels of a given analyte plus three times the standard deviation.

134 Concentration data were non-normal despite log transformation and therefore treated non-
135 parametrically for statistical analyses; summary statistics and group comparisons were derived
136 using uncensored data analyzed using the cenfit function in the R package NADA version 1.6 -
137 1.1 to account for artifacts of left-censored data^{44,45}. Significant differences in contaminant
138 concentrations between individuals or other ecologically relevant groups were assessed using both
139 uncensored and censored log-transformed data. Left-censored data was also assessed for
140 significant differences by group and compound using Kruskal-Wallis tests followed by post-hoc
141 application of Dunn's test for multiple comparisons. Data is presented as median \pm SD.

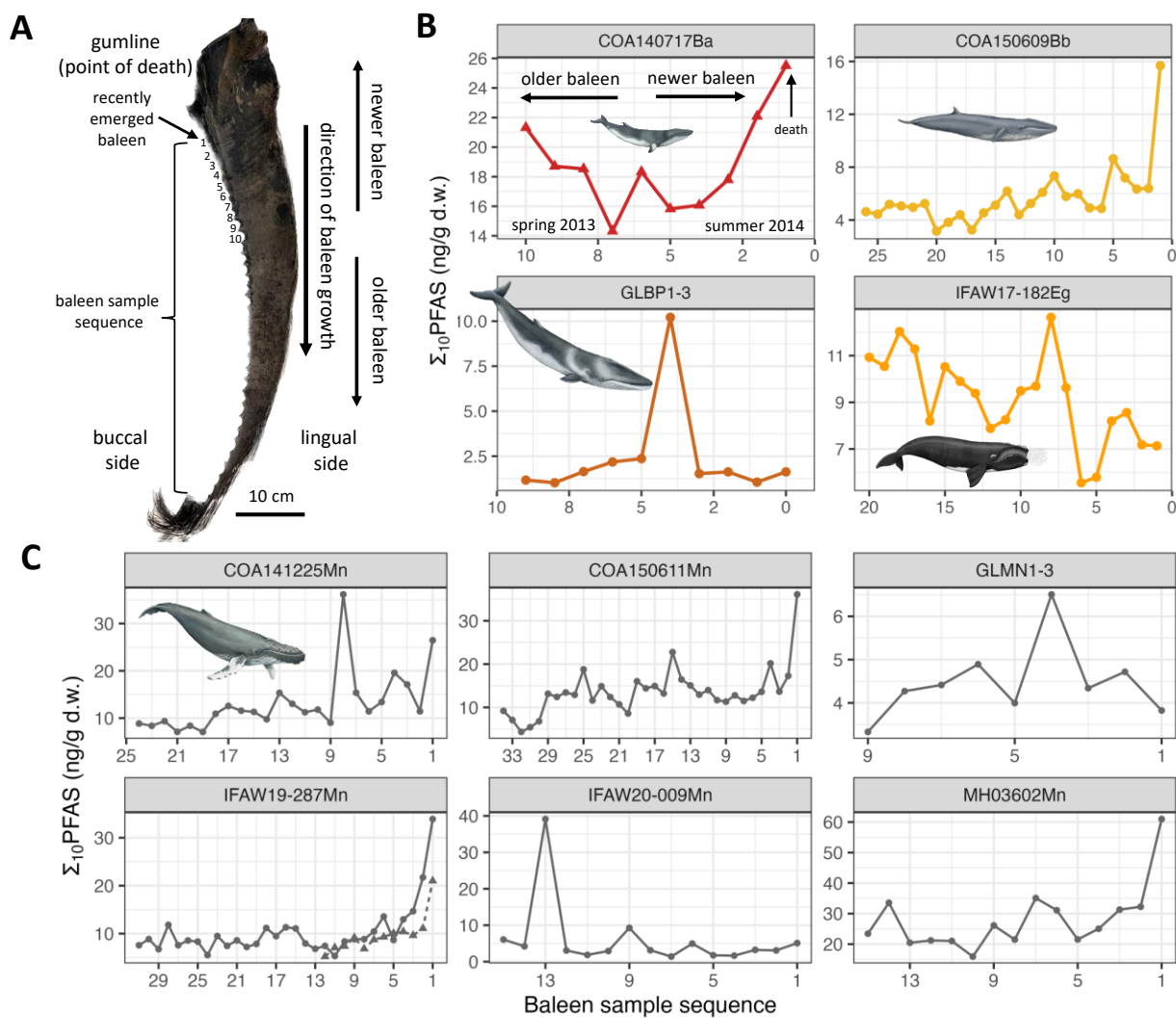
142

143 **Results and Discussion**

144 *PFAS concentrations in baleen*

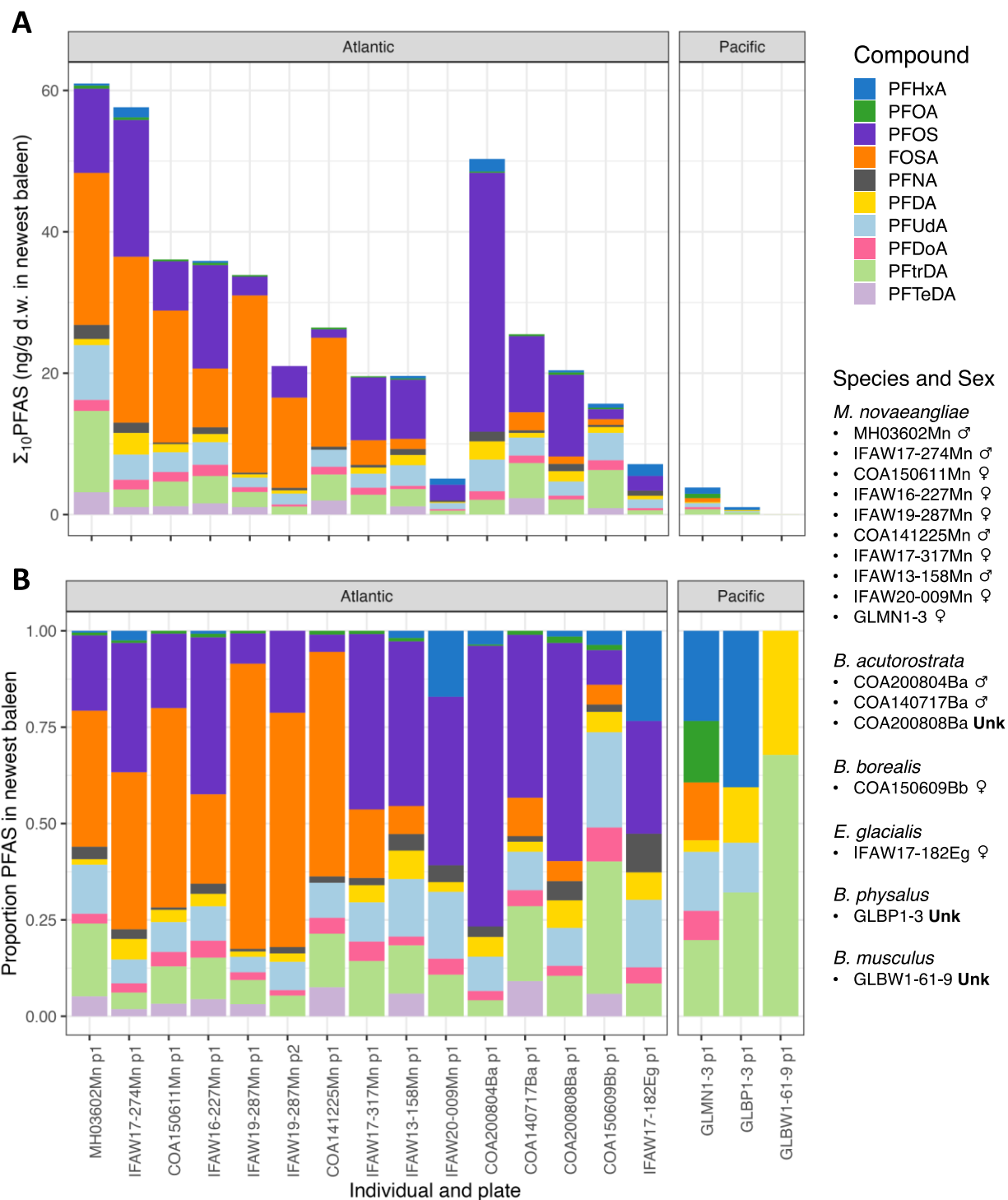
145 We found quantifiable levels of PFAS in every baleen sample (n = 220) across 18 plates
146 from six species (Fig. 1, Table S14). Across all tissue types we tested, 16 of 26 targeted PFAS
147 compounds were quantifiable in at least one sample (for more information on these compounds,
148 see the Supplemental Information). In baleen samples, ten PFAS – PFUdA, PFTTrDA, PFDA,
149 PFDoA, PFOS, PFOA, PFNA, FOSA, PFHxA, and PFTeDA, referenced as Σ_{10} PFAS hereafter –
150 – were found in > 40% of samples. The median Σ_{10} PFAS for all baleen samples was 10 ± 10 ng/g
151 dry weight (d.w.). Using data from the most recently emerged baleen (i.e., the first ~2 cm segment
152 on each plate that was fully erupted from the gumline) on each plate, the Σ_{10} PFAS ranged from
153 0.02 ng/g d.w. for a blue whale from the Pacific Coast to 61 ng/g d.w. for a humpback whale from
154 the Atlantic Coast (Fig. 2). The PFAS concentrations we report here represent lower-bound
155 estimates of the total PFAS-burden in these whales, as there are likely additional PFAS compounds
156 that we did not detect with our evaluation of 26 targeted analytes¹⁵. Of the 26 target analytes, half

157 were below instrumental detection limits in baleen samples (4:2-FTS, 6:2-FTS, 8:2-FTS,
 158 EtFOSAA, 7:3 FTCA, HFPODA-GenX, N-MeFOSAA, Nafion BP2, PFBA, PFNS, PFHpS,
 159 PFPeA, and PFPeS), but there were quantifiable levels of PFUdA, PFTTrDA, PFDA, and PFDoA
 160 – all long-chain perfluoroalkyl carboxylic acids (PFCAs) – in > 80% of baleen samples. The
 161 abundance of these long-chain PFCAs has been noted in other marine food webs, potentially driven
 162 by long-range transport and transformation of PFCA intermediates to remote ocean and polar
 163 regions^{46–48}. Of the 10 compounds found in > 40% of baleen samples, PFOS – a long-chain
 164 perfluoroalkyl sulfonic acid (PFSA) – had the highest median concentration 2 ± 4 ng/g d.w.



165
 166 **Fig. 1. PFAS load along baleen plates.** Data showing sampling points and PFAS data for samples

167 taken 2-3cm along each plate. Ten PFAS compounds appeared in > 40% of all baleen samples
168 along baleen plates, the Σ_{10} PFAS of these compounds are plotted. Longitudinal data for Common
169 minke, sei, fin, and blue whale not shown. A) Baleen plate sampled from a humpback whale
170 (IFAW19-287Mn) showing sampling locations, direction of growth, and orientation in the oral
171 cavity. B) Common minke (n = 3 total; one shown here), sei, fin, and North Atlantic right whales;
172 blue whale not shown due to short plate available for analysis and thus no time series. C) humpback
173 whales plotted by individual, only individuals with > 3 samples per plate are shown; note that for
174 IFAW19-287 two plates were sampled one represented with circular points and solid line, another
175 with triangular points and dashed line. Whale illustrations by Alex Boersma.



176
 177 **Fig. 2. PFAS in recently emerged baleen.** P1 refers to the first plate sampled for each individual,
 178 p2 refers to the second plate sampled. Only IFAW19-287Mn had two separate plates sampled.
 179 Refer to Table S1 for more information on each individual A) concentration of the Σ_{10} PFAS for
 180 the subsample of most recently emerged baleen on each plate studied colored by compound. B)
 181 proportion of PFAS compounds for the most recently emerged baleen on each plate studied
 182 colored by compound.

183

184 *Trends along a baleen plate*

185 In addition to the universal detection of PFAS in our baleen samples, there were trends in
186 PFAS contamination along each plate that may be related to life history events. Most commonly,
187 there was a spike in Σ_{10} PFAS in the final 3-4 samples prior to death (Fig. 1). This could be due to
188 changes in habitat use, behavior, feeding, or body condition in the months prior to death. However,
189 the yearling North Atlantic right whale (IFAW17-182Eg; NARW Catalogue #4694) baleen we
190 tested showed the opposite pattern, with a decreasing PFAS-baleen burden over time (Fig. 3A).
191 This may be due to rapid weight gain of the individual as it grew from newborn to yearling; typical
192 NARW birthweight is 1 ton and its final weight, 7.5 tons, was measured at necropsy⁴⁹ (Fig. 3A).
193 Baleen whales have determinate growth—growing rapidly in their first two years before somatic
194 growth slows as they approach sexual maturity (5-10 years old in most species)^{49,50}. In addition,
195 during their first year, baleen whales transition from feeding on milk at birth to prey as a yearling
196 – NARWs fully wean by 8-17 months⁵¹. It is likely that mammalian milk is enriched in PFAS as
197 compared to zooplankton. Based on baleen growth rates and the known age of this NARW at its
198 time of death, the steadily decreasing Σ_{10} PFAS in its baleen likely reflects body dilution in the
199 growing calf that occurred as its diet shifted from milk to copepods (Fig 3A). Furthermore, the
200 necropsy report indicated that the cause of death was probable blunt trauma from a vessel strike.
201 An acute cause of death (e.g., vessel strike) reduces the possibility of other factors (e.g., chronic
202 stress, starvation, or a change in habitat use following injury or entanglement in fishing gear)
203 causing long-term changes in body chemistry that can be reflected on a baleen plate^{35,36,52}.

204 In contrast, IFAW19-287Mn, “Vector,” was a mature female humpback whale that had
205 been studied for 35 years at the time of her death and had a documented calving history. Based on

206 a baleen growth rate of 12 cm/yr for an adult humpback²⁵ and her time of death, the length of the
 207 longer of her two plates we sampled (67 cm, representing ~5 years of growth) contained the
 208 gestation, parturition, and nursing periods of her final two calves. Long-chain PFASs (e.g., PFOS)
 209 are known to transfer from mother to offspring in the womb and through milk⁵³. The portion of
 210 baleen that coincides with these two reproductive periods shows two declines in PFOS levels,
 211 which may represent maternal offloading *in utero* and via milk (Fig. 3B). In the future, larger
 212 sample sizes of females with known reproductive histories and concurrent hormonal analysis of
 213 baleen would provide further opportunities to examine maternal offloading of various PFAS
 214 compounds.

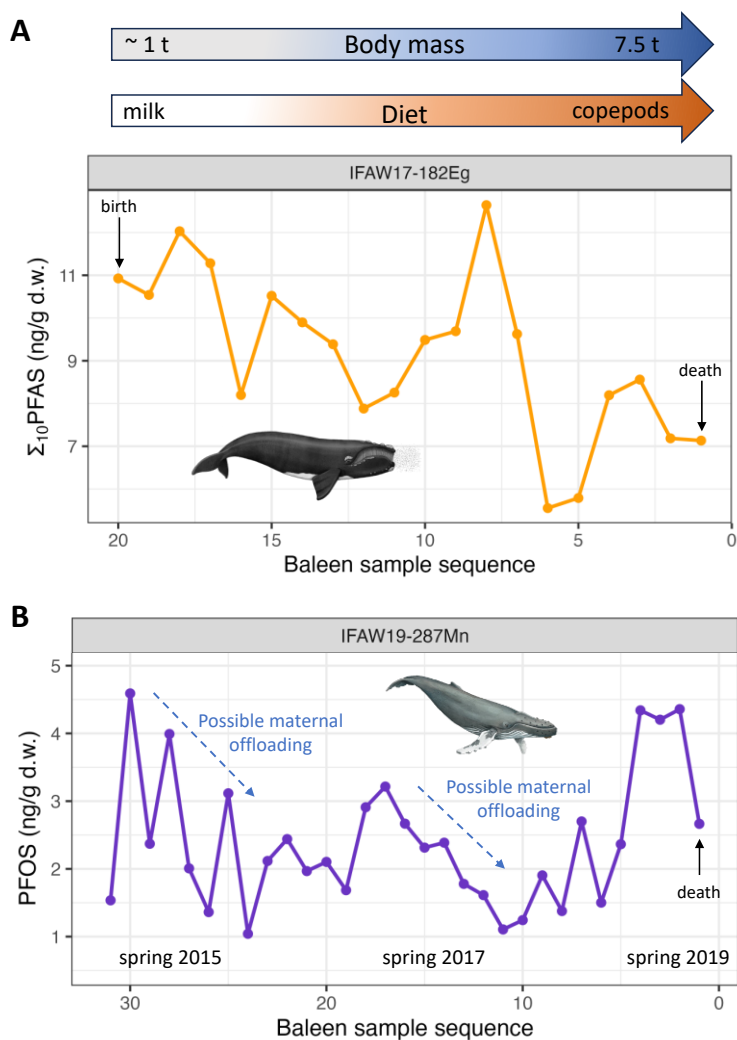


Fig. 3. Baleen-PFAS signature and life-history events. A) Longitudinal record of NARW baleen-PFAS signal with samples every ~3 cm along the plate; the only baleen plate in our study that captured the entire life of the individual. This individual was 15 months old at its time of death (likely via ship strike) and capturing the weaning period. B) The longitudinal PFOS record of the longer of Vector's (IFAW19-287Mn) two plates that were analyzed, showing declining PFOS-levels coinciding with two known calving periods, one in winter-spring 2015 the other in winter-spring 2017. Whale illustrations by Alex Boersma.

239

240 *Geographic trends*

241 We did not have adequate sample size or geographic coverage to statistically assign
242 differences in PFAS values between whales from the Atlantic vs. Pacific coasts of North America.
243 However, our initial results suggest higher PFAS concentrations in marine food webs in the Gulf
244 of Maine compared to the California Current (Fig. 2A). This is in agreement with available global
245 data and models on PFAS distributions in seawater^{47,64}. The median Σ_{10} PFAS for the most recently
246 emerged baleen samples was 26 ± 17 ng/g d.w. for individuals from the Gulf of Maine ($n = 15$),
247 as compared to 1 ± 3 ng/g d.w. for individuals from the U.S. West Coast ($n = 3$). Preliminary
248 differences this large suggest the need for further investigation. Globally, it would be useful to
249 compare our findings – across species and compounds – with data from the highly polluted
250 Mediterranean and Western North Pacific regions where data on PFAS and large whales are
251 limited and outdated.

252

253 *PFAS across whale tissues*

254 Of the five tissues we tested, only liver ($n = 5$, 4 humpbacks, 1 NARW) had consistently
255 higher PFAS-burdens than baleen. The average PFAS burden in liver samples was fivefold higher
256 than what was found in baleen, but roughly twentyfold higher than that in blubber and skin (Fig.
257 S2, Table S15). PFAS concentrations by tissue type were liver > baleen \approx gum > skin > blubber
258 (Fig. S2). Skin and blubber are the only tissues that can be reliably sampled from living cetaceans
259 due to restrictions on interactions with marine mammals mandated by global and US law; however,
260 our findings suggest these are the least effective tissues for PFAS monitoring across the five tissue
261 types we tested. This is in contrast with lipophilic POPs that are well represented in blubber^{54–56}.

262 While our wet tissues sample size was limited, our findings suggest tissue-specific partitioning of
263 PFAS. For instance, 7:3 FTCA was found in all liver samples (n = 5), 38% of blubber samples (3
264 out of 8), 25% of skin samples (2 out of 8) but was absent from all gum and baleen samples (n =
265 228 total). In contrast, median PFOS values in baleen were tenfold higher than PFOS in blubber
266 (2.1 ng/g vs. 0.24 ng/g). Thus, multi-tissue sampling provides a more complete contaminant
267 picture than any single tissue alone, providing rationale for future efforts to assess PFAS in
268 multiple tissues from marine mammals^{16,57}.

269

270 *Comparison to other marine mammals*

271 Most studies on PFAS in marine mammals have tested liver tissue. We used our liver
272 samples to compare to other findings for PFAS in marine mammals. In our study, we found eight
273 PFAS that were present in > 40% of wet tissue (liver, skin, blubber) samples that we included in
274 our analyses. Those compounds were PFUdA, PFNA, PFDA, PFOS, FOSA, PFtrDA, 7:3 FTCA,
275 and PFDoA. Our Σ_8 PFAS concentrations for liver samples were 75 ± 85 ng/g wet weight (w.w.),
276 which is lower than the majority of pinniped seal and odontocete whale studies^{15,16,18,19}, but aligns
277 well with other mysticete whale data (Table S16). This was expected as baleen whales tend to have
278 lower trophic positions than pinnipeds or odontocetes⁵⁸. These levels are within range to cause
279 immune suppression in mice (64 - 118 ng/g w.w. serum PFOS)⁵⁹ and humans (6 - 21 ng/mL serum
280 PFOS)^{60,61}; no comparable toxicological testing has yet been conducted in any marine mammal.

281 As with other cetacean studies^{62,63}, we recorded FOSA in all liver samples (16 ± 14 ng/g
282 w.w.) and in most baleen samples (61% of baleen samples; 1 ± 4 ng/g d.w.). Humpback whales
283 had significantly higher concentrations of FOSA than the other species in our dataset, which may
284 be indicative of their nearshore residency and feeding (humpback whale recently emerged baleen

285 FOSA: 13 ± 10 ng/g d.w.; other species' baleen FOSA: 0 ± 1 ng/g d.w.; Kruskal-Wallis test, $P =$
286 0.007). In addition, we found 7:3 FTCA in higher concentrations than previously published
287 mysticete studies (Table S16), though recent work uncovered high levels of 7:3 FTCA in killer
288 whales (*Orcinus orca*) off the North American west coast¹⁹. It is likely that air-breathing species
289 have higher levels of 7:3 FTCA as compared to fish and aquatic invertebrates due to atmospheric
290 oxidation of fluorotelomer alcohols as the likely origin of environmental 7:3 FTCA.

291

292 *Implications*

293 We report quantifiable levels of PFAS in all samples ($n = 251$) across five tissue types in
294 six species of baleen whales from the Atlantic and Pacific coasts of North America with
295 concentrations spanning five orders of magnitude (0.02 ng/g Σ_{10} PFAS in blue whale baleen to >
296 200 ng/g Σ_8 PFAS in a humpback whale liver). While PFAS have been found in other keratin-based
297 tissues, including bird feathers and human hair^{10,65-67}, this is the first report of these contaminants
298 in baleen. Future work should explore whether there are additional PFAS present in baleen. Our
299 initial work focused on a list of common PFAS compounds, while other research has indicated that
300 there might be additional organofluorine present in marine mammals, including baleen whales¹⁵.

301 The proof-of concept work presented here relied on fairly large baleen samples (~0.25g /
302 sample) to ensure detection of the targeted PFAS present. A more precise temporal resolution in
303 the PFAS chronology could be achieved by using lower masses, and thus increase the temporal
304 resolution. The size, ease of storage, longitudinal growth, and demonstrated PFAS binding affinity
305 makes baleen a valuable resource for monitoring PFAS in the environment. Moreover, museum
306 baleen collections dating back to the 19th-century can provide a window into the global emergence
307 of PFAS in marine ecosystems across space and time.

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551 **Data and materials availability:**
552 All data are available in the main text or the Supporting Information; Code for figures and
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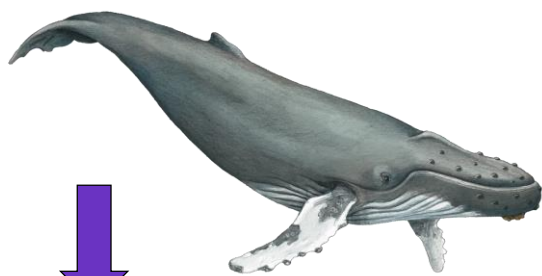
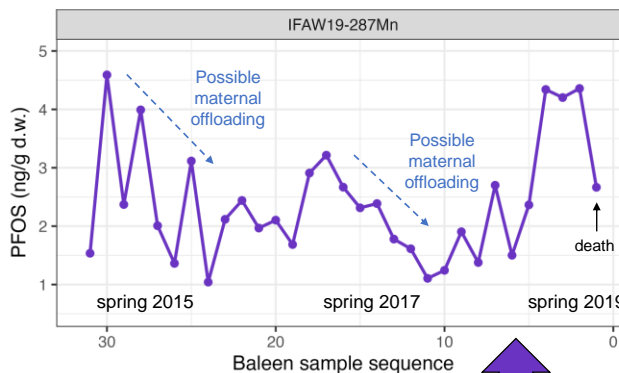


Image: Center for Coastal Studies under NMFS permit 18786



0.7 meters