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## A new, global optical sediment trap calibration

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## A new, global optical sediment trap calibration

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## A new, global optical sediment trap calibration

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### Abstract

Autonomous sensors for gravitational carbon flux in the ocean are critically needed, because of uncertainties in the projected response of the biological carbon pump (BCP) to climate change, and the proposed, engineered acceleration of the BCP to sequester carbon dioxide in the ocean. Optical sediment trap (OST) sensors directly sense fluxes of sinking particles in a manner that is independent of, and complementary to, other autonomous, sensor-derived estimates of BCP fluxes. However, limited intercalibrations of OSTs with traditional sediment traps and uncharacterized, potential biases have limited their broad adoption. A global field data set spanning three orders of magnitude in carbon flux was compiled and used to develop empirical models predicting particulate organic carbon flux from OST observations, and intercalibrating different sensor designs. These data provided valuable constraints on the uncertainty in the predicted carbon flux and showed a quantitative, theoretically consistent relationship between observations from OSTs with collimated and diffuse optical geometries. While not designed for this purpose, commercial beam transmissometers have been used as OSTs, so two models were developed quantifying the biases arising from the transmissometer's housing geometry and optical beam diameter. Finally, an algorithm for the quality control of beam transmissometer-derived OST data was optimized using sensitivity tests. The results of this study support the expansion of OST-based gravitational carbon flux measurements and provide a framework for interpretation of OST measurements alongside other gravitational particle flux observations. These findings also suggest key features that should be included in designs of future, purpose-built OST sensors.

The ocean's biological carbon pump (BCP) plays a key role in regulating Earth's climate by transferring biologically derived carbon to the ocean's interior, thereby setting the vertical dissolved inorganic carbon (DIC) gradient (Volk and Hoffert 1985). The gravitational carbon pump (GCP), in which particulate organic carbon (POC) passively sinks away from the euphotic zone, is thought to make up about half of the total BCP flux globally

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Additional Supporting Information may be found in the online version of this article.

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(Boyd et al. 2019). Sinking particles in the ocean form through a variety of biological and physical processes which are closely tied to the surface ocean ecosystem state (De La Rocha and Passow 2007; Siegel et al. 2016). Changes in the efficiency of the BCP could lead to climate feedbacks that are poorly constrained at present by Earth system models (Henson et al. 2022). Limited observations of biological carbon fluxes in the global ocean present a key obstacle to improved predictions. In particular, model improvements require observations of biological carbon flux that differentiate key mechanisms of the GCP, and these observations have traditionally required labor-intensive, ship-based methods. In addition, the responsible implementation of marine carbon dioxide removal through nutrient addition to the ocean (e.g., as proposed by National Academies of Sciences, Engineering, and Medicine 2022) will require the rapid deployment of accurate GCP flux sensors that are able to operate independently of ships.

A growing network of autonomous ocean-observing platforms now provides estimates of total BCP fluxes through geochemical budgets of DIC, nitrate, and oxygen, but these estimates do not distinguish the contribution of the GCP (Nicholson et al. 2008; Johnson et al. 2017; Huang et al. 2022). Profiling platforms that

measure optical backscattering ( $b_{bp}$ ) provide POC flux estimates by measuring change in the depth-integrated POC inventory vs. time (Dall'Olmo and Mork 2014), and are calibrated using a broad global data set of POC :  $b_{bp}$  measurements (e.g., Cetinić et al. 2012). This method cannot distinguish sinking from suspended particles under steady-state conditions, nor can it constrain fluxes below the profile depth, but the simplicity and robustness of optical backscattering sensors permit its broad application (Briggs et al. 2011; Alkire et al. 2012; Dall'Olmo and Mork 2014; Bol et al. 2018; Lacour et al. 2019; Briggs et al. 2020; Giering et al. 2020a).

Occasionally, profiling floats equipped with beam transmissometers (Bishop et al. 2004; Bishop and Wood 2009; Estapa et al. 2013, 2017, 2019b) or upward-looking cameras (Bishop et al. 2016; Bourne et al. 2019, 2021) are used as optical sediment traps (OSTs), providing proxy measurements that are specific to gravitational POC flux and can be used under steady-state conditions. In the OST method, sinking particles are intercepted by the sensor itself (the upward-facing optical window of a vertically mounted transmissometer or camera) and the rate of change in the measured light attenuation provides the proxy for flux. However, the direct intercalibration of an OST with a co-deployed, traditional sediment trap is labor-intensive and only two published data sets are available, for the subtropical North Atlantic Ocean (Estapa et al. 2017) and the California Coastal Current (Bourne et al. 2019).

Here, we comprehensively evaluate the performance of the OST method and provide a global calibration, with a goal of broadening the user base of the method, as the need for well-calibrated POC flux measurements grows. We compile a global set of paired attenuation and sediment trap POC flux data from multiple sources spanning contrasting ocean biogeochemical settings, from the oligotrophic subtropics to the North Atlantic spring bloom. We describe an updated algorithm used for retrieval of attenuation flux from a timeseries of OST measurements, and then present attenuation data that have been derived from two sensor types: in situ beam transmissometers, which measure the attenuation of collimated light using a non-imaging sensor as described above; and laboratory photomicrographs of polyacrylamide gel sediment traps (Durkin et al. 2021) from which a mean estimate of uncollimated, diffuse attenuation can be derived (described below). We compare these data to published, diffuse attenuation and POC flux observations described by Bourne et al. (2019), and by Huffard et al. (2020). Taken together, these data provide POC flux calibrations for OST-like beam attenuation and diffuse attenuation sensors, that can be extrapolated broadly in the global ocean. We explore the uncertainties associated with the use of different OST sensor types for POC flux measurements, and finally, we discuss the implications of the relationship between beam attenuation and diffuse attenuation for the design of improved sinking particle flux sensors.

## Materials and procedures

### Measurement principles

The term attenuation (atn) is used here to refer to the negative log-transformed ratio of the light passing through a plane to a detector ( $I_d$ ), to the light incident upon that plane ( $I_o$ ; Eq. 1; both have units of  $W m^{-2} nm^{-1}$ ).

$$atn = -\ln\left(\frac{I_d}{I_o}\right) \quad (m^2 m^{-2}) \quad (1)$$

Note that atn as used here follows the definition of Bishop et al. (2016) and is analogous to absorbance (Abs) as defined by Mobley (2022), who instead defined *attenuance* as  $I_d/I_o$ . If the light source is collimated and the detector acceptance angle is close to  $0^\circ$  (subject to practical limitations in reality; Boss et al. 2009) then we specify that atn is the beam attenuation ( $atn_c$ ), which is related to the more familiar beam attenuation coefficient ( $c$ ) of light passing through a sample volume over the pathlength  $L$  (Eqs. 2, 3).

$$c = -\frac{\ln(I_d/I_o)}{L} \quad (m^{-1}) \quad (2)$$

$$atn_c = c \times L \quad (m^2 m^{-2}) \quad (3)$$

where the sampled area of  $atn_c$  and the sampled volume of  $c$  are related by  $L$ . While atn is a unitless quantity and  $c$  has units of  $m^{-1}$ , it is helpful to think of both as ratios, in the case of atn between the sample's total attenuation cross section ( $\sigma_c$ , units of  $m^2$ ; Mobley 2022) and the sensor's illuminated area ( $m^2$ ), and in the case of  $c$  between  $\sigma_c$  and the volume illuminated by the sensor ( $m^3$ ). Thus, atn has units of ( $m^2 m^{-2}$ ), where the numerator refers to the attenuation cross section of the light-attenuating material and the denominator to the geometric area. The attenuation cross section of a single particle can be thought of as the particle's cross sectional area multiplied by the summed fractions of incident light absorbed and scattered by that particle. A particle that is completely opaque, and large relative to the wavelength of the light, will have an attenuation cross section that is twice the geometric cross section of that particle (Mobley 2022).

It is also useful to define a second variant of atn for the situation where the illuminating light source is uncollimated and the detector has a wide acceptance angle. This is the case if atn is measured using a camera (Bishop et al. 2016; Bourne et al. 2019, 2021). Here, we adopt the name  $atn_{K_d}$  which is analogous to the downward, diffuse attenuation coefficient  $K_d$  ( $K_d(z) = -d/dz \ln(E_d(z))$ , where  $E_d(z)$  is the depth-resolved plane irradiance; Mobley 2022). Similar to the relationship between  $c$  and  $K_d$  in the water column (Mobley 2022), where the incident light field becomes less diffuse and the sensor acceptance angle decreases (for  $c$  relative to  $K_d$ ), for the same sample we expect  $atn_c$  to always be greater than  $atn_{K_d}$  because the former excludes more of the scattered

light from the detector. As the incident light field becomes less and less diffuse, the value of  $atn_{K_d}$  is expected to approach that of  $atn_c$ .

The rate of change of  $atn$  with respect to time serves as the proxy for particle flux ( $F_{atn}$ ; Eq. 4).

$$F_{atn} = \frac{datn}{dt} \left( m^2 m^{-2} d^{-1} \right) \quad (4)$$

The proxy assumes that  $atn$  is only changing due to the accumulation of sinking particles, and not due to changes along the optical path of the transmissometer such as the transient passage of zooplankton, changes in the background optical properties due to vertical motion of the sensor through the water column, or changes in the background pixel brightness across an image. Therefore, some pre-processing of  $atn$  data is usually required prior to estimation of  $F_{atn}$ . Below, we describe the general methods we have developed to remove interferences from two types of  $atn$  data: (1) timeseries of  $atn_c$  collected using beam transmissometers used as OSTs on autonomous floats drifting at a sampling depth and (2)  $atn_{K_d}$  computed from images of particles accumulated in sediment traps.

#### Processing $atn_c$ data from in situ transmissometer as OST

A beam transmissometer (C-Rover 2000, 650 nm, 0.25 m pathlength, Sea-Bird Scientific) deployed as an OST on a drifting autonomous platform returned measurements of sensor depth ( $z$ ), time ( $t$ ), and  $atn_c$  at intervals of 3, 5, or 15 min, collected over drift periods of 2–6 d (see “Data sets” section for descriptions of different data sets presented here). If a rapid sequence of  $atn_c$  measurements was collected at each timepoint (i.e., “burst sampling”), these were averaged and timepoints with a standard deviation greater than  $0.0013 m^{-2} m^2$  were discarded. Points collected outside of  $\pm 25$  m from the targeted depth were also discarded. Spikes, which Estapa et al. (2017) observed to be well correlated with zooplankton swimmer intrusion to nearby sediment traps, were removed from the data using a moving median filter followed by a moving mean filter (Briggs et al. 2011) with a 1.5-h window. As described by Estapa et al. (2017), the despiked data typically contained periods with slowly increasing  $atn_c$  interrupted by rapid increases, and sometimes decreases, in  $atn_c$  over a single sampling interval. While we interpret rapid increases as sinking particle flux (see “Effect of sampling cross section” and “Effect of optical sampling cross section” sections), rapid decreases can only be explained by loss of accumulated material from the sampling surface, possibly due to turbulence or zooplankton consumption (Estapa et al. 2017). Losses of  $atn_c$  at a rate more negative than  $-0.05 m^2 m^{-2} d^{-1}$  were therefore cumulatively removed from the data series. After these corrections, the data series were smoothed with a centered, moving mean filter with a 1.5 h window, and the first difference with respect to time computed to obtain  $F_{atn_c}$  (Eq. 4), which includes both the rapid

and slowly increasing periods of the  $atn_c$  timeseries. Finally, a single mean  $F_{atn_c}$  value was averaged for the whole drift period for comparison to simultaneously collected sediment trap POC fluxes (see “Data sets” section). A Matlab (Mathworks) script carrying out the above steps is also provided (Supplemental Text; Estapa 2023).

#### Processing $atn_{K_d}$ data from sediment trap imagery

The majority of the images of sinking particles reported in this study were obtained through laboratory microscopy of polyacrylamide gel collectors that were deployed in cylindrical sediment traps as described by Durkin et al. (2021). Here we describe the image processing procedure to derive a mean value of  $atn_{K_d}$  for each gel. Similar image processing procedures were followed to obtain  $atn_{K_d}$  from the in situ “Sedimentation Event Sensor” (SES; McGill et al. 2016) camera data set described in “California Coastal (Sta. M)” section. Briefly, polyacrylamide gel is a clear, viscous fluid placed in the bottom of a sediment trap, that preserves sinking particle morphology until samples can be imaged under a microscope in the laboratory. Here, the gel samples were imaged with a color digital camera (various models) under transmitted light with eight-bit resolution, at low magnification with a pixel size ranging from 7.9 to 15  $\mu m$  (see “Data sets” section). The red channel of the color images was used for all analyses for best correspondence to the 650-nm C-Rover wavelength. A set of 10–20, non-overlapping image tiles was collected for each sample. A value of  $atn_{K_d}$  was computed for each image pixel of each tile. However, unlike measurements of  $atn_c$  described above, multiple image timepoints were not available from which to compute  $datn/dt$ . Instead, we assumed that prior to particle accumulation,  $I_d = I_o$  for each image pixel (that is, at time = 0,  $atn_{K_d} = 0 m^2 m^{-2}$ ). To obtain a value of  $I_o$  for each pixel, we assumed that the background lighting characteristics stayed the same across the  $N$  images collected with the same camera and microscope setup (i.e., on the same cruise). If undeployed “blank” gel images were collected, these were used to provide the appropriate pixel-specific  $I_o$  values. If no blank gel images were available, then a composite background image was created from the sample images (see Supplement Figure SF1 for an illustration). Assuming a sufficiently large number of images with randomly located particles, the background pixel value ( $I_o[i,j]$ ) was set equal to the brightest value for that pixel over the whole set of  $N$  images (Eq. 5).

$$I_o(i,j) = \max_N I_d(i,j,N) \quad (5)$$

Prior to calculating  $atn_{K_d}$ , each image tile was visually examined for artifacts such as bubbles, zooplankton that were likely to have actively entered the trap, and sample jar edges. Masks were created manually to eliminate these areas from further analysis. Then,  $atn_{K_d}$  images were calculated (Eq. 1) and an area-weighted mean  $atn_{K_d}$  value was computed over all



tiles for each sample. Finally, the collection period of the sediment trap deployment was used as the time interval to compute  $F_{\text{atn}_{K_d}}$ .

### Evaluation of beam transmissometers as OSTs

While beam transmissometers have seen the widest use as OSTs due to their commercial availability and ease of integration to autonomous sampling platforms, they were not designed for the purpose of intercepting and detecting relatively large sinking particles. We tested the degree of bias likely to arise through use of a transmissometer as an OST due to imperfect filtering of environmental artifacts and instrument noise, self-shading by the instrument housing, and undersampling of rare, large particles by an optical beam with a small cross section.

#### *Sensitivity tests of algorithm for determining $F_{\text{atn}_c}$ from a beam transmissometer*

While the basis for determination of the  $\text{atn}_c$  rate of change is straightforward, the process of filtering out instrumental noise and environmental artifacts, like spikes and rapid negative excursions in  $\text{atn}_c$ , has the potential to introduce bias. We performed a series of sensitivity tests on an artificial data set to evaluate the performance of our algorithm for retrieving  $F_{\text{atn}_c}$  from a beam transmissometer. Simulated data from a transmissometer OST were generated with different characteristics including  $F_{\text{atn}_c}$  magnitude ( $0.01\text{--}0.69\text{ m}^2\text{ m}^{-2}\text{ d}^{-1}$ ), constant vs. variable  $F_{\text{atn}_c}$ , presence/absence of positive and negative excursions, and different levels of random instrument noise. These simulated data characteristics were chosen to span the ranges that are typically observed in real deployments. The parameters of the  $F_{\text{atn}_c}$  algorithm (the despiking and smoothing filter window width, and the threshold for identifying “negative jumps”) and the overall averaging period were systematically varied to find the settings that resulted in the lowest root mean squared error (RMSE) (between “true” and retrieved  $F_{\text{atn}_c}$ ) over the largest range of input data characteristics. The parameters tested are shown in Table 1.

#### *Self-shading*

The design of a transmissometer includes upper and lower pressure housings that respectively contain the source and detector optics. When vertically mounted as for OST use, the upper housing is above the detector window, so particles must be

carried into the sampling volume by ambient turbulence in the water. Under quiescent conditions or with rapidly sinking particles, the upper housing is most likely to preclude particle collection, which we refer to here as “self-shading” (Estapa et al. 2017). Using a rough turbulence scaling estimate, Estapa et al. (2017) estimated that at a dissipation rate of  $\epsilon = 10^{-9}\text{ m}^2\text{ s}^{-3}$  (a midrange value for the upper 1 km; Waterhouse et al. 2014) self-shading would begin to impact particles settling faster than  $200\text{ m d}^{-1}$ , with greater effects on faster settling particles and at lower  $\epsilon$  values. Estapa et al. (2017) concluded that the impact of self-shading was likely negligible in that study which sampled the upper 500 m in a subtropical gyre where POC flux was dominated by small, slowly settling particles (Durkin et al. 2015). In the present study, we now reconsider the impact of transmissometer self-shading in more productive settings with faster-sinking particles, and at deeper depths with less turbulence.

To test the extent of self-shading, we created a simple, particle tracking model on top of the output from a publicly available, non-dimensional large-eddy simulation (LES) of a turbulent flow field (Li et al. 2008). We scaled the modeled, non-dimensional velocities output from the LES to a range of dissipation rates and kinematic viscosities ( $\nu$ ) that are characteristic of the ocean interior ( $\epsilon = 10^{-11}\text{--}10^{-7}\text{ m}^2\text{ s}^{-3}$  and  $\nu = 0.87\text{--}1.6 \times 10^{-6}\text{ m}^2\text{ s}^{-1}$ ; Waterhouse et al. 2014). Twenty simulations with different  $\epsilon$  and  $\nu$  values were completed. Simulated particles (1000 per  $\epsilon\text{--}\nu$  combination) were randomly assigned sinking speeds from a lognormal distribution ranging from 1 to  $1000\text{ m d}^{-1}$ . These were released at random times and locations in the model and followed through time to generate sinking particle “tracks.” Then, we generated a polyhedron representation of a C-Rover transmissometer (10 cm diameter) adjacent to a profiling float hull (15 cm diameter), and translated the particle tracks so that they all terminated within the transmissometer’s beam cross section at the detector window. Using the MATLAB Central File Exchange function “inpolyhedron.m” (Sven 2023) we detected which simulated particle tracks intersected the transmissometer housing, and thus would have been affected by “self-shading.” The fraction of shaded particles and the 2.5<sup>th</sup>, 25<sup>th</sup>, 50<sup>th</sup>, 75<sup>th</sup>, and 97.5<sup>th</sup> percentiles of their sinking speed distributions were tabulated and compared to the sinking speed distribution of the initial particle population.

#### *Effect of sampling cross section*

The light beam in most commercial beam transmissometers has a relatively small cross section, generally on the order of 1 cm diameter where it intersects the detector window on which sinking particles accumulate. This is only 1–2 orders of magnitude larger than the sinking particles themselves, which have diameters on the order  $100\text{ }\mu\text{m}$  to 1 mm in size. Thus, a transmissometer beam cross section is small enough that long averaging periods may be necessary to adequately characterize the true flux. We characterized the tradeoff between averaging period and sampling cross section by analyzing an image timeseries collected using an in situ camera sensor looking

**Table 1.** Parameters tested in  $F_{\text{atn}_c}$  retrieval algorithm.

Filter window (h)	Negative jump threshold ( $\text{m}^2\text{ m}^{-2}\text{ d}^{-1}$ )	Overall averaging period (d)
1	0	0.25
1.5	−0.005	0.5
2	−0.05	1
4	−0.5	2

Parameter values shown in boldface are the optimized values.

upwards through a polyacrylamide gel sediment trap. Particles larger than 300  $\mu\text{m}$  were identified and tracked in the image timeseries, and then the “area flux” of particles (particle area per time) within subsampled areas of varying diameter (0.7–5 cm) was calculated and compared to the flux in the entire image (see “Effect of optical sampling cross section” section).

### Data sets

To assess the empirical relationships among  $F_{\text{atn}_c}$ ,  $F_{\text{atn}_{K_d}}$ , and POC flux ( $F_{\text{POC}}$ ), we compiled all available data sets that included  $F_{\text{atn}_c}$ ,  $F_{\text{atn}_{K_d}}$ , or both, as well as co-located  $F_{\text{POC}}$  from sediment trap measurements. Table 2 summarizes the key information chronologically, and the subsections under the “Data sets” section briefly describe the details and deviations of each deployment. The mean geographic locations where the data were collected are summarized in Fig. 1. As much as possible, we refer the reader to the original publications of these data sets for deployment, sample analysis details, and links to the original data, in order to minimize repeated information. Where new data are presented here for the first time, we provide an expanded description of these details. The individual measurements are given in Supplemental Table ST1.

#### Subtropical North Atlantic (Bermuda Atlantic Timeseries Study)

This site is an oligotrophic, subtropical gyre setting characterized by a low productivity, highly recycling ecosystem. Data were collected monthly from July to September (stratified period) and also just prior to the spring bloom (March). Sediment trap deployments and sample analyses are described by Estapa et al. (2017). The deployment depths spanned from 150 to 500 m. Key caveats are that the particulate carbon flux is from particles 350  $\mu\text{m}$  and smaller due to the screens used to remove zooplankton “swimmers,” and total particulate carbon rather than POC

was analyzed. However, passively sinking particles at this site were small, and the particulate inorganic carbon contribution to the flux tends to be negligible (Durkin et al. 2015; Estapa et al. 2017). Sediment trap samples were wet-split in triplicate prior to analysis and we take the standard deviation among splits as the uncertainty in  $F_{\text{POC}}$  (Fig. 2). Although polyacrylamide gel traps were deployed (Durkin et al. 2015), they were not imaged under transmitted light, so no  $\text{atn}_{K_d}$  estimates are available.

#### California coastal (Sta. M)

This site underlies a productive, upwelling eastern boundary current system. Data were collected over the course of 8 months (October 2014–June 2015; Huffard et al. 2020). Measurements were made at a depth of 3900 m, with POC fluxes collected in a moored, conical trap (Parflux, McLane, Inc.) and images of sinking particles measured in situ with the transmitted-light, imaging SES located 3.2 km from the sediment trap. The deployments, trap sample analyses, and SES instrument are described elsewhere (McGill et al. 2016; Huffard et al. 2020). The SES images (10.5281/zenodo.7613681) were reanalyzed here to provide estimates of  $F_{\text{atn}_{K_d}}$  using the “composite background” method described in “Processing  $\text{atn}_{K_d}$  data from sediment trap imagery” section. In addition to POC fluxes, mass fluxes were also available from this deployment. Estimated uncertainties in  $F_{\text{POC}}$  are not available.

#### New England coastal

This site is a productive near-coastal setting, with measurements collected in November (fall diatom bloom) and June (early summer, with flux dominated by zooplankton fecal matter). Sediment traps were deployed at a depth of 150 m, seaward of the shelf break current. These deployments and sample collections are described and links to the data are provided by Durkin et al. (2021). Sediment trap tubes were deployed in triplicate and we take the standard deviation

**Table 2.** Data set summary.

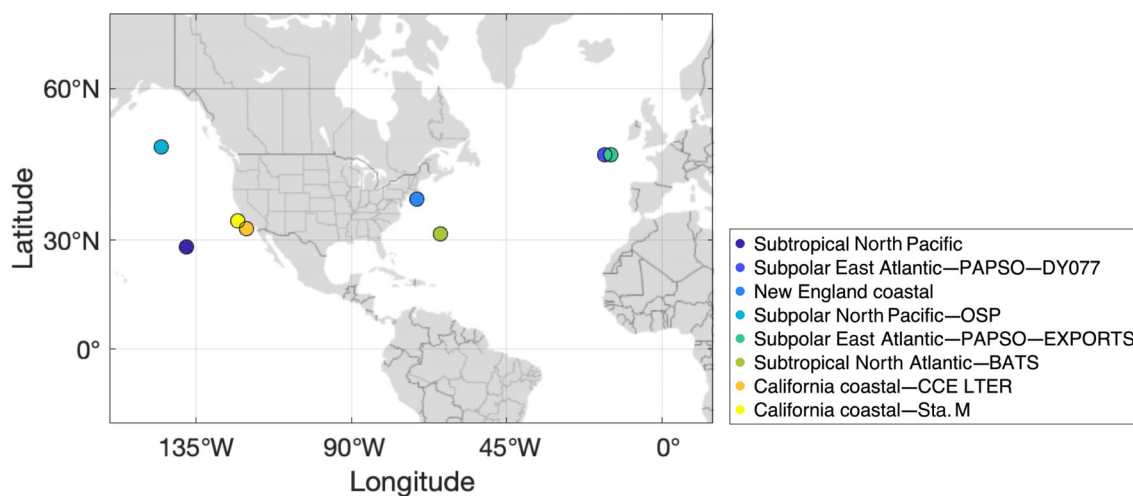
Region (program)	Year	References*	Position (degrees)	$F_{\text{atn}_c}$ method	$F_{\text{atn}_{K_d}}$ method
Subtropical North Atlantic (BATS)	2013–2014	1	31.67N–64.17E	C-Rover on NBST	Gel trap microscopy
California coastal (Sta. M)	2014–2015	2, 3	34.83N–123.00E	None	Sedimentation Event Sensor†
New England coastal	2015–2016	4, 5	39.80N–70.97E	C-Rover on NBST	Gel trap microscopy
Subtropical North Pacific	2017	5, 6	28.25N–137.68E	C-Rover on NBST	Gel trap microscopy
Subpolar Northeast Atlantic (PAPSO)	2017	4, 7	48.98N–16.48E	C-Rover on NBST	Gel trap microscopy
California coastal (CCE LTER)	2017	8, 9	32.87N–120.28E	None	Carbon Flux Explorer‡
Subpolar North Pacific (Sta. P-EXPORTS)	2018	6, 10	50.45N–144.99E	C-Rover on NBST	Gel trap microscopy
Subpolar Northeast Atlantic (PAPSO-EXPORTS)	2021	6	48.92N–14.79E	C-Rover on NBST	Gel trap microscopy

BATS, Bermuda Atlantic Timeseries Study; CCE LTER, California Current Ecosystem Long-Term Ecological Research; EXPORTS, Export Processes in the Ocean Through Remote Sensing; PAPSO, Porcupine Abyssal Plain Sustained Observatory.

\*References: (1) Estapa et al. (2017); (2) Huffard et al. (2020); (3) McGill et al. (2016); (4) Estapa et al. (2019b); (5) Durkin et al. (2021); (6) *this study*; (7) Baker et al. (2020); (8) Bourne et al. (2019); (9) Bishop et al. (2016); (10) Estapa et al. (2021).

†Described in “California Coastal (Sta. M)” section.

‡Described in “California Coastal (California Current Ecosystem Long-Term Ecological Research)” section.



**Fig. 1.** Locations of sample collection sites summarized in Table 2. Colors correspond to those in Fig. 2.

among tubes as the uncertainty in  $F_{\text{POC}}$  (Fig. 2). Gel sediment traps were imaged at a resolution of  $15 \mu\text{m pixel}^{-1}$ .

#### **Subtropical North Pacific**

Sediment trap deployments were carried out at a depth of 150 m during two process stations in the subtropical North Pacific in wintertime, as part of the “Sea2Space” project (as described by Durkin et al. 2021). One station was representative of a low-productivity subtropical gyre setting, and the other was located in the productive California coastal current. Sediment trap samples were wet-split in triplicate prior to analysis and we take the standard deviation among splits as the uncertainty in  $F_{\text{POC}}$  (Fig. 2). The new  $F_{\text{atn}_{K_d}}$  estimates were derived from the transmitted light images of these samples (resolution  $7.9 \mu\text{m pixel}^{-1}$ ) published by Durkin et al. 2021.

#### **Subpolar Northeast Atlantic (Porcupine Abyssal Plain Sustained Observatory)**

Sediment trap deployments were carried out at depths of 200 and 350 m during a sediment trap intercomparison experiment conducted at the Porcupine Abyssal Plain Sustained Observatory (PAPSO) site in the subpolar Northeast Atlantic, just prior to the spring bloom. The deployments and sample analyses are described by Baker et al. (2020). Sediment trap samples were wet-split in triplicate prior to analysis and we take the standard deviation among splits as the uncertainty in  $F_{\text{POC}}$  (Fig. 2). The transmitted-light microscopy on the gels was collected with a pixel size of  $11.5 \mu\text{m}$ . In addition to POC fluxes, mass fluxes were also available from this project. Data from this project are available from Estapa et al. (2019a).

#### **California coastal (California Current Ecosystem Long-Term Ecological Research)**

The measurements from this study were collected in the productive, upwelling California coastal current system, and descriptions of particle characteristics were given by Bourne et al. (2021). The POC and  $\text{atn}_{K_d}$  flux data from this study are

reported without modification from the original publication (Bourne et al. 2019), except for a change of logarithmic base to match the definition of  $\text{atn}_{K_d}$  used here. Observations were collected at approximately 150 m with an autonomous, drifting, imaging sediment trap (“Carbon Flux Explorer,” Bishop et al. 2016). The in situ camera sensor aboard the CFE looked upward into the base of a conical sediment trap, and collected series of images over 25-min time intervals, interspersed with cleaning cycles prior to collection of background images. Unlike the other  $\text{atn}_{K_d}$  estimates reported here, a threshold was applied prior to taking the image average, so that pixels with  $\text{atn}_{K_d} < 0.046$  were set to zero (Bourne et al. 2019). Uncertainties in  $F_{\text{POC}}$  were propagated from the standard deviations of replicate determinations of POC from each trap sample (Bourne et al. 2019).

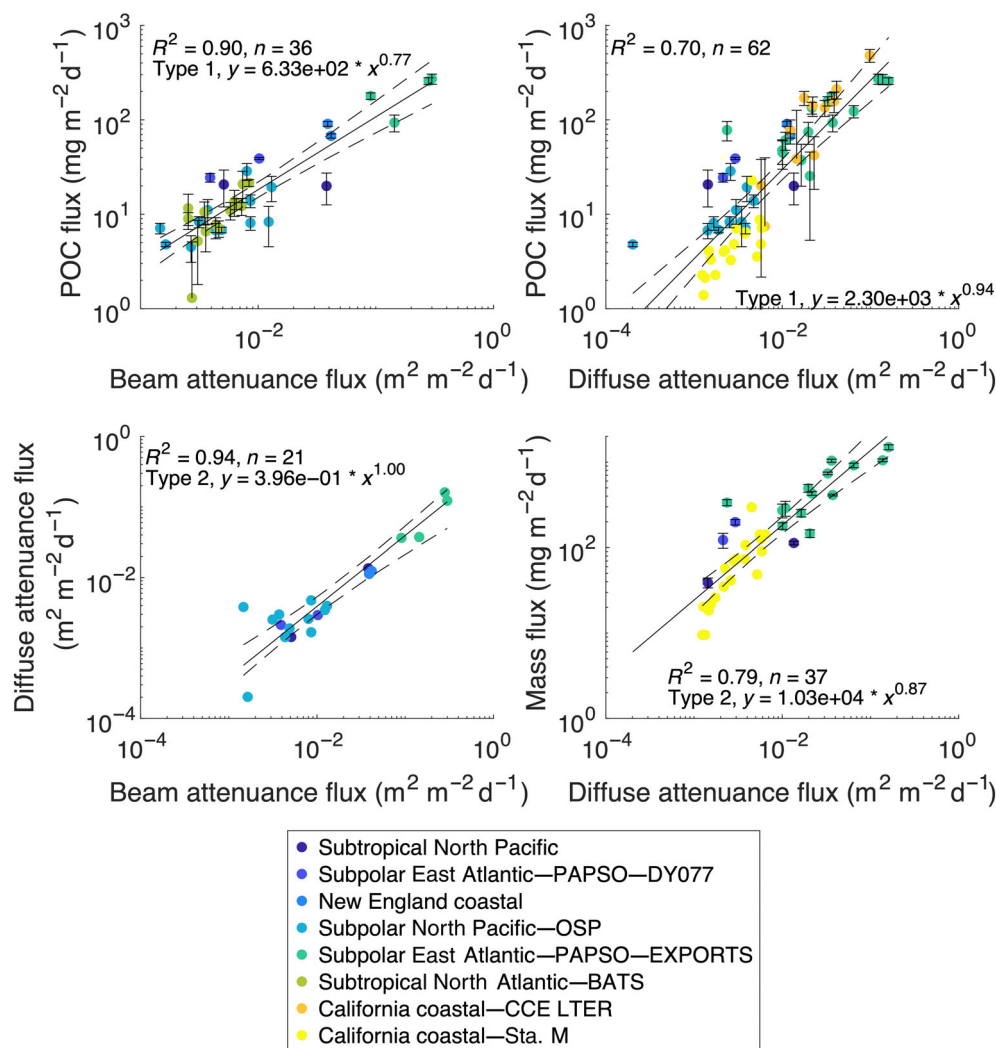
#### **Subpolar North Pacific (Sta. P, Export Processes in the Ocean Through Remote Sensing)**

These observations were collected in a low-productivity, iron-limited “high nutrient low chlorophyll” subpolar gyre, with the low particle flux dominated by zooplankton products and episodic salp blooms (Durkin et al. 2021; Steinberg et al. 2023). The sediment trap deployments and sample analyses from this study are described by Estapa et al. (2021), and the data are available in Estapa and Durkin (2018). Sediment trap samples were wet-split in triplicate prior to analysis and we take the standard deviation among splits as the uncertainty in  $F_{\text{POC}}$  (Fig. 2). The gel traps were imaged at a resolution of  $7.9 \mu\text{m pixel}^{-1}$ . In this study, the sediment trap samples were subject to a heavy contamination by zooplankton “swimmers,” which was difficult to eliminate during sample processing. A post-analysis, statistical correction was also performed (Estapa et al. 2021).

#### **Subpolar North Atlantic (PAPSO, Export Processes in the Ocean Through Remote Sensing)**

The data from this study, whose trap deployments captured a high-flux event associated with the termination of the North





**Fig. 2.**  $F_{POC}$  and  $F_{atn}$  flux models (lines) compared to observational data (colored points). Vertical error bars show uncertainty in  $F_{POC}$  observations, obtained as described in the subsections under the “Data sets” section. No horizontal error bars are shown because  $F_{atn}$  can be determined very precisely, while biases must be constrained through modeling approaches described in “Evaluation of beam transmissometer performance as an OST” section. Solid line shows model fit to data and dashed lines show the 95% confidence interval of the fit. Best fit equations,  $R^2$  values, and number of observations ( $n$ ) are also given. Colors correspond to Fig. 1. (Upper left) Prediction of  $F_{POC}$  from  $F_{atn}$ . (Upper right) Prediction of  $F_{POC}$  from  $F_{atn_{kd}}$  (lower left). Relationship between  $F_{atn_{kd}}$  and  $F_{atn}$ . (Lower right) Prediction of  $F_{mass}$  from  $F_{atn_{kd}}$ .

Atlantic spring diatom bloom, have not been previously published. Briefly, POC fluxes were measured with a combination of neutrally buoyant sediment traps (NBSTs; Estapa et al. 2020) and surface-tethered sediment traps. All traps carried polyacrylamide gels deployed and analyzed similarly to Durkin et al. (2021), while NBSTs also carried beam transmissometers (e.g., C-Rover 2000) used as OSTs as described by Estapa et al. (2020). Gels were imaged at a resolution of  $7.9 \mu\text{m pixel}^{-1}$ . The sediment traps were deployed in three cycles, for 2–6 d periods at depths from 75 to 500 m in the center of an anticyclonic eddy near the PAPSO site during the Export Processes in the Ocean Through Remote Sensing (EXPORTS) North Atlantic field campaign (Johnson et al., in

prep). The sampling procedures are described briefly here, but correspond closely to the more detailed methods described by Estapa et al. (2021), with the exception that a statistical correction for zooplankton swimmer contamination was not required. Trap tubes were cylinders with collection areas of  $0.0113 \text{ m}^2$  (12 cm diameter). Two tubes on each trap were used for POC flux measurements and were filled with 500 mL of 0.1% formaldehyde-poisoned, 70 ppt brine overlain with filtered seawater. One tube on each trap carried a polyacrylamide gel collector overlain by filtered seawater. Upon retrieval, the two brine tubes were combined, samples were passed through 350- $\mu\text{m}$  screens, the screens picked clean of zooplankton, and then the screen contents were washed back into the  $<350 \mu\text{m}$  sample.

Samples were then split on a custom rotary splitter (Lamborg et al. 2008). Splits for POC were filtered onto precombusted GF/F filters (Whatman), and dried until analysis. The GF/F filters were split gravimetrically and portions of three filters from each sample were analyzed for total C via combustion elemental analysis, and inorganic C via coulometric titration. POC flux was computed as the difference between the total and inorganic fluxes. We take the standard deviation among splits as the uncertainty in  $F_{\text{POC}}$  (Fig. 2). Sample splits for mass flux were filtered onto pre-treated polycarbonate membranes then dried and reweighed, until stable to within 0.01 mg, to determine the mass flux. Data from this study are available in Estapa and Durkin (2022).

### Models relating POC flux to attenuation flux

We tested several empirical models to predict POC flux and mass flux from  $F_{\text{atn}_c}$  and  $F_{\text{atn}_{K_d}}$ , and to characterize the relationship between  $\text{atn}_c$  and  $\text{atn}_{K_d}$ . To first order, we expected a linear Beer's law relationship to apply, meaning we expected light attenuation to be proportional to the concentration of the attenuating substance. However the compiled data set (Supplemental Table ST1) included fluxes spanning three orders of magnitude, with different particle types from biogeochemically contrasting settings, so the data were log-transformed prior to carrying out regression analyses (Eq. 6). Where errors in the  $x$ - and  $y$ -variables were comparable (e.g., between  $\text{atn}_c$  and  $\text{atn}_{K_d}$ ) we also tested Type II regression models, otherwise we used Type I models.

$$F_{\text{POC}} = 10^{a_2} \times (F_{\text{atn}})^{a_1} \left( \text{mg-C m}^{-2} \text{d}^{-1} \right) \quad (6)$$

The performances of the different empirical models were evaluated through the RMSE and normalized RMSE (nRMSE) between the observations and the model, and through their correlation coefficients. We also ruled out significant non-linearity of the models by testing for the potential significance of a  $(\log[F_{\text{atn}}])^2$  term.

### Assessment

#### Models predicting POC flux from attenuation flux

The global data set supports the prediction of POC flux from  $F_{\text{atn}_c}$  or image-mean  $F_{\text{atn}_{K_d}}$ , and the prediction of mass

flux from  $F_{\text{atn}_{K_d}}$ . The regression coefficients, RMSE, and nRMSE values for the models are provided in Table 3 and the model fits are plotted with the observations in Fig. 2. None of the models were significantly nonlinear ( $p < 0.05$ ). There was more scatter in the  $F_{\text{atn}_{K_d}} - F_{\text{POC}}$  relationship compared to  $F_{\text{atn}_c}$  ( $R^2 = 0.90$  vs.  $0.70$ , Table 3), but the  $F_{\text{atn}_{K_d}} - F_{\text{POC}}$  relationship showed more linearity than did  $F_{\text{atn}_c} - F_{\text{POC}}$  ( $a_1 = 0.94 \pm 0.18$  vs.  $0.77 \pm 0.14$ ; Table 3; Fig. 2). The strongly linear relationship between  $F_{\text{atn}_c}$  and  $F_{\text{atn}_{K_d}}$  ( $a_1 = 1.00 \pm 0.09$ ;  $R^2 = 0.94$ ; Table 3; Fig. 2) is consistent with theoretical expectations (see "Diffuse attenuation flux from transmitted-light, imaging sensors" section). The performance of  $F_{\text{atn}_{K_d}}$  as a proxy for mass flux was stronger than that of  $F_{\text{atn}_c}$  (RMSE = 199 vs. 291  $\text{mg m}^{-2} \text{d}^{-1}$ ).

#### Evaluation of beam transmissometer performance as an OST

The good correlation between  $F_{\text{atn}_c}$  and  $F_{\text{POC}}$  suggests that biases due to algorithm assumptions, self-shading, and undersampling were at least consistent across the roughly three order-of-magnitude flux data set analyzed here. The possible impacts of these biases were quantified through a series of modeling and sensitivity tests described below.

#### Sensitivity tests of the $F_{\text{atn}_c}$ data processing algorithm

The six simulated data sets, and the retrieval of  $F_{\text{atn}_c}$  from them using the algorithm described in "Processing  $\text{atn}_c$  data from in situ transmissometer as OST" section and by Estapa et al. (2017), are depicted graphically in Fig. 3. Two constant, mean  $\delta \text{atn}_c / \delta t$  magnitudes were simulated, with and without spikes and jumps. A variable  $\delta \text{atn}_c / \delta t$  at an intermediate magnitude was also simulated with and without spikes and jumps. Together, the six simulated data sets covered a  $\delta \text{atn}_c / \delta t$  range of  $0.01 - 0.69 \text{ m}^2 \text{ m}^{-2} \text{ d}^{-1}$ . For each of the six simulated data sets, five different instrument noise levels were simulated, corresponding to noise standard deviations (SD) ranging from  $0 \times 10^{-3}$  to  $8.6 \times 10^{-3} \text{ m}^2 \text{ m}^{-2}$  (Supplement Table ST2a). As a point of comparison, the manufacturer-reported precision of the 25-cm pathlength C-Rover 2000 used to collect the  $F_{\text{atn}_c}$  data discussed here is  $c = 0.003 \text{ m}^{-1}$ , equivalent to  $\text{atn}_c = 7.5 \times 10^{-4} \text{ m}^2 \text{ m}^{-2}$ . In most cases when noise SDs were equal to or smaller than  $8.6 \times 10^{-4} \text{ m}^2 \text{ m}^{-2}$ , the algorithm was

**Table 3.** Summary of models tested, with coefficients as in Eq. 6.

$x$	$y$	Regression type	$N$	$R^2$	$a_1$ (unc)	$a_2$ (unc)	RMSE (nRMSE)	Units for $y$ and RMSE
$F_{\text{atn}_c}$	$F_{\text{POC}}$	1	36	0.90	0.77 (0.14)	2.8 (0.3)	19.38 (0.52)	$\text{mg-C m}^{-2} \text{d}^{-1}$
$F_{\text{atn}_{K_d}}$	$F_{\text{POC}}$	1	62	0.70	0.94 (0.18)	3.4 (0.4)	49.76 (0.81)	$\text{mg-C m}^{-2} \text{d}^{-1}$
$F_{\text{atn}_c}$	$F_{\text{atn}_{K_d}}$	1	21	0.94	0.93 (0.18)	-0.54 (0.36)	0.01729 (0.85)	$\text{m}^2 \text{m}^{-2} \text{d}^{-1}$
$F_{\text{atn}_c}$	$F_{\text{atn}_{K_d}}$	2	21	0.94	1.00 (0.09)	-0.40 (0.18)	0.01183 (0.58)	$\text{m}^2 \text{m}^{-2} \text{d}^{-1}$
$F_{\text{atn}_c}$	$F_{\text{mass}}$	1	7	0.77	0.65 (0.47)	3.4 (0.8)	291 (0.59)	$\text{mg m}^{-2} \text{d}^{-1}$
$F_{\text{atn}_{K_d}}$	$F_{\text{mass}}$	1	37	0.79	0.87 (0.16)	4.0 (0.4)	199 (0.74)	$\text{mg m}^{-2} \text{d}^{-1}$

able to recover the original  $\delta\text{atn}_c/\delta t$  to within 10% over the tested range and in the presence of large positive and negative “jumps” in the signal (Supplement Table ST2b,c; noise SD of  $8.6 \times 10^{-4} \text{ m}^2 \text{ m}^{-2}$  case illustrated in Fig. 3). However, in the simulations with the highest noise level and the lowest  $\delta\text{atn}_c/\delta t$ , the true signal was not recoverable (Fig. 4; Supplement Table ST2b,c). The algorithm’s success at retrieving  $\delta\text{atn}_c/\delta t$  was not especially sensitive to the negative jump threshold or to smoothing window widths  $\leq 2\text{ d}$ , but was more sensitive to the relative instrument noise (Supplement Table ST2b,c). This suggests that the most appropriate parameter choices will take into account the sensor-platform combination and how the OST is sampled and logged by the platform controller. For the simulated data sets used here, which mimicked the OST data collected by C-Rovers carried by NBSTs (Estapa et al. 2020), the optimal averaging window length was 1.5 h and the optimal negative jump threshold was  $-0.05 \text{ m}^2 \text{ m}^{-2} \text{ d}^{-1}$ .

### Results of self-shading analysis

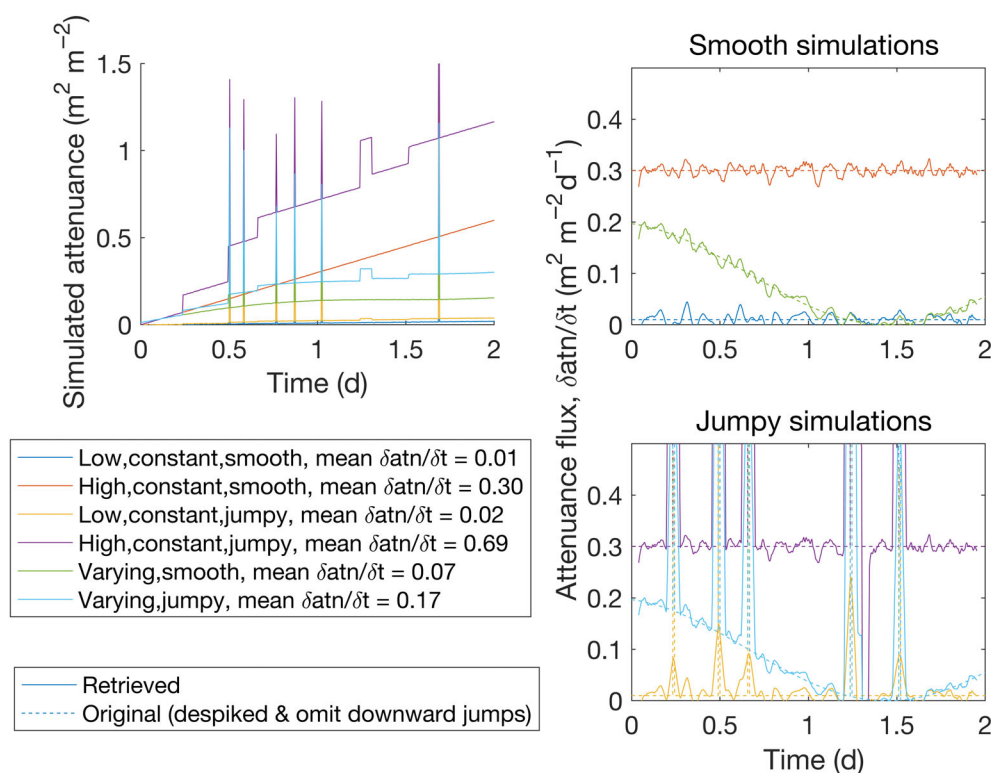
An example of sinking particle tracks simulated by the self-shading model is illustrated in Fig. 5. The turbulent dissipation rate ( $\epsilon$ ) had a much greater effect on the extent of particle shading than did the viscosity (Fig. 6) so the results are

summarized here as a function of  $\epsilon$  only. As expected, there was more shading under lower-dissipation rate conditions, with 20–25% of particles blocked at  $\epsilon = 10^{-11} \text{ m}^2 \text{ s}^{-3}$  but less than 5% blocked at  $\epsilon \geq 10^{-8} \text{ m}^2 \text{ s}^{-3}$  (Fig. 6).

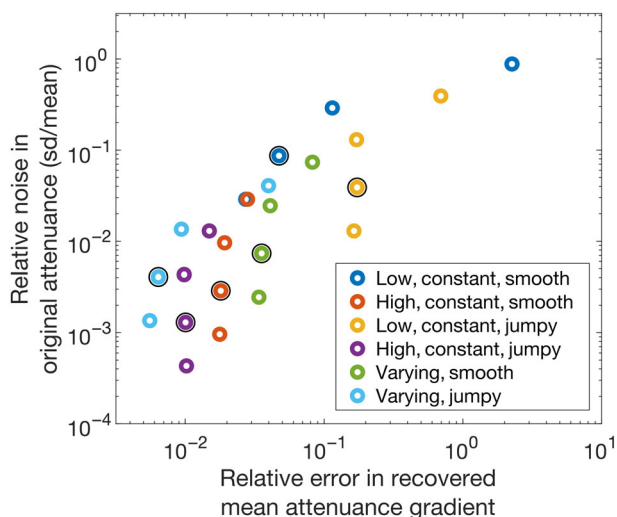
Also as expected, particles with faster sinking speeds were disproportionately shaded compared to slower-sinking particles. We compared the sinking speed distribution of the shaded particles to that of the whole modeled population and found that the shaded particles had higher sinking speeds (Fig. 7). For instance, the median sinking speed of the modeled particles was  $29 \text{ m d}^{-1}$ , while the median sinking speed of shaded particles ranged from 45 to  $558 \text{ m d}^{-1}$ . The sinking-speed dependence of shading was most pronounced at lower  $\epsilon$ , whereas at the largest modeled value ( $\epsilon = 10^{-7} \text{ m}^2 \text{ s}^{-3}$ ), the distribution of shaded sinking speeds began to approach that of the initial population (Fig. 7).

### Effect of optical sampling cross section

In situ, time-lapse video imagery from a gel trap sample in the subtropical North Pacific data set (“Subtropical North Pacific” section) allowed us to examine the choice of “beam” or imaged area on attenuation flux. The camera, mounted beneath the transparent bottom of the gel cup, took an image every 20 min, for a duration of 61 h. Each 5 megapixel



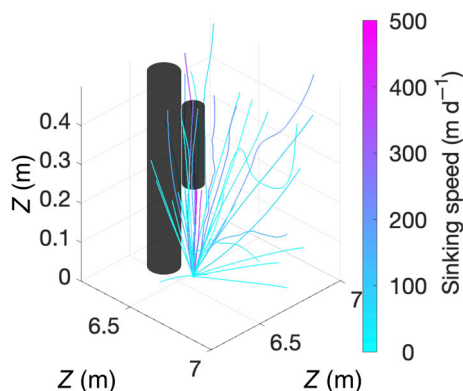
**Fig. 3.** Simulated  $\text{atn}_c$  timeseries and ability of algorithm (“Processing  $\text{atn}_c$  data from in situ transmissometer as OST” section) to retrieve the  $\text{atn}_c$  rate of change ( $\delta\text{atn}_c/\delta t$ ). (Upper left panel) Original, simulated data sets with overall mean  $\delta\text{atn}_c/\delta t$  values ( $\text{m}^2 \text{ m}^{-2} \text{ d}^{-1}$ ) shown in the legend. (Upper right panel) Plots of  $\delta\text{atn}_c/\delta t$  vs. time for simulated data sets without spikes and jumps (dashed lines) and the  $\delta\text{atn}_c/\delta t$  retrieved using the algorithm described in “Processing  $\text{atn}_c$  data from in situ transmissometer as OST” section (solid lines). (Lower right panel) Same as upper right panel, only for the three data sets that included spikes and jumps. Timeseries shown here have added instrument noise with standard deviation of  $8.6 \times 10^{-4} \text{ m}^2 \text{ m}^{-2}$ .



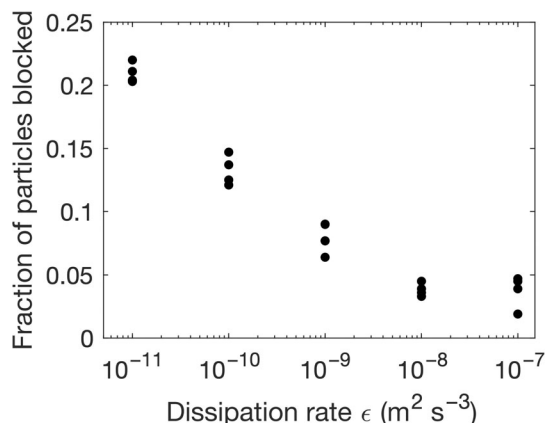
**Fig. 4.** Relationship between the retrieved, mean  $\delta \text{atn}_c / \delta t$  and the amount of instrument noise in the original attenuation signal (tested levels were 1, 3, 10, and  $30 \times 10^{-3} \text{ m}^2 \text{ m}^{-2}$ ). Relative noise is expressed as the standard deviation of the noise divided by the mean attenuation in the original simulated data. Symbol colors and short legend labels correspond to the simulated data sets shown in Fig. 3, upper left panel. Symbols outlined in black correspond to the timeseries shown in Fig. 3 (noise SD of  $0.003 \text{ m}^2 \text{ m}^{-2}$ ).

image was cropped to remove blurring and distortion, resulting in a 6.8 cm by 5.6 cm imaged area and a pixel size of approximately  $50 \mu\text{m}$  (Fig. 8). The smallest size of particle that could be resolved from the images had an equivalent spherical diameter of about  $300 \mu\text{m}$ .

The fractional cross-sectional area occluded by particles was calculated for the entire imaged area (thick black line, Fig. 8a–c), resulting in a timeseries that resembled the simulated



**Fig. 5.** Illustration of model used to test extent of transmissometer self-shading. The large and small gray cylinders are the hulls of a profiling float and the upper housing of a transmissometer, respectively. Trajectories of sinking particles (colored lines; color denotes speed) were modeled at random locations in the turbulent velocity field and then translated to intersect the transmissometer’s detector window. A random subsampling of 50 trajectories (out of 1000 total) are shown.



**Fig. 6.** Fraction of particles blocked by the transmissometer or float hull as a function of dissipation rate. Scatter among points at each  $\epsilon$  value represents the dependence upon kinematic viscosity (itself primarily a function of temperature), and was much smaller than the dependence on  $\epsilon$ .

attenuance shown in Fig. 2. Next, each image was subsampled with different sizes of “beam diameter,” with the smallest diameter ( $\sim 0.7 \text{ cm}$ ) corresponding to the beam of a C-Rover transmissometer (Fig. 8a). A Monte Carlo sampling approach was applied by randomly selecting the central position of each subsampled region, and generating a timeseries of fractional cross-sectional area covered by particles. These experiments were repeated 10 times for each of the subsampled “beams.” The means of these experiments (thin black lines, Fig. 8a–c) and the standard deviations (gray shaded areas) show an expected increasing trend over time, with significantly larger stochasticity in the experiments with small beam diameter. The root mean square deviation (RMSD, Eq. 7; Fig. 8d–f) was calculated from

$$\text{RMSD}^2 = \frac{\sum_{i=1}^N (\text{atn}_i - \text{atn}_{\text{true}})^2}{N} \quad (7)$$

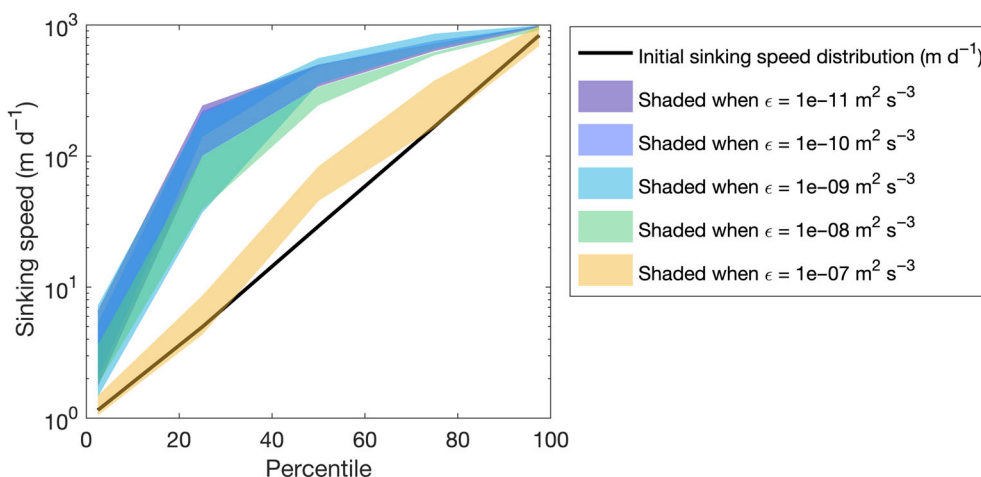
and showed that as the simulated beam diameter increases, the results converge toward the case with the full imaged area (of  $38 \text{ cm}^2$ ). Here, we found that  $> 90\%$  of the total variability was captured by a beam area greater than  $5 \text{ cm}^2$ . The results highlight that the discrete arrivals of large particles have a strong impact on the time variability of the area flux (and also of  $F_{\text{atn}}$ ), particularly when the subsampled area is small.

## Discussion

### Attenuance flux as a quantitative proxy for POC flux

The importance of the ocean’s BCP to global climate (Volk and Hoffert 1985), and its centrality to proposals to artificially enhance ocean biological carbon uptake (National Academies of Sciences, Engineering, and Medicine 2022), demand that we expand global observations of the GCP in the highly undersampled waters below the mixed layer. To that end,



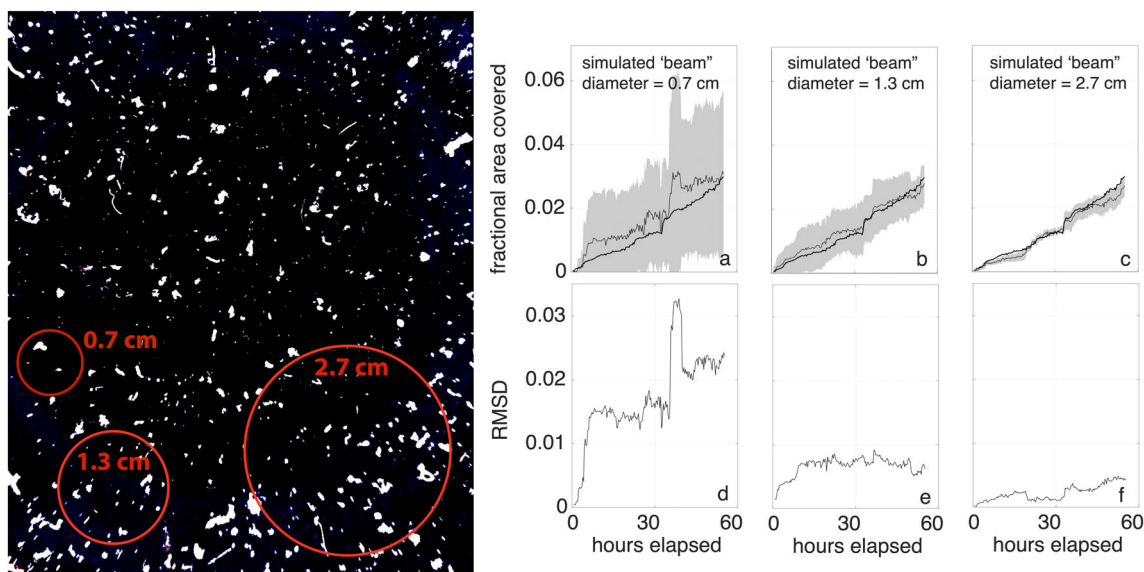


**Fig. 7.** Comparison of sinking speed distributions of particles that were blocked by the transmissometer or float hull (colored, shaded patches) to the speed distribution of the initial particle population (black line). The turbulent dissipation rate  $\epsilon$  is indicated by color. The vertical breadth of the colored areas is caused by differences due to kinematic viscosity and random variability (since the number of blocked particles was small).

proxies that can accurately quantify GCP fluxes from autonomous platforms, and without ship support, will be especially valuable (Claustre et al. 2021). Attenuance flux can be estimated through a variety of methods, all of which measure the light transmitted through accumulating, sinking particles. Many of these methods have been, or can be adapted for autonomous use.

Multiple prior studies have observed a correspondence between POC flux and atn flux in regional settings (Bishop et al. 2004, 2016; Bishop and Wood 2009; Estapa et al. 2013,

2017, 2019b; Bourne et al. 2019, 2021). This expected, quantitative relationship is analogous to Beer’s Law which predicts that light attenuation will follow mass concentration in a water volume, although in the present application, attenuation and mass accumulation are measured in an area rather than a volume. Here, we have confirmed this expectation, and shown that global, empirical models can respectively describe the  $POC : atn_c$  and  $POC : atn_{K_d}$  relationships across three orders of magnitude in flux and across contrasting ocean biogeochemical settings. These settings have sinking particles of varying



**Fig. 8.** Effect of area sampled by an optical detector on flux signal variability. (Left panel) An example of a processed image, with  $> 300 \mu\text{m}$  particles shown in white. Red circles indicate the sizes of sub-sampled areas. (Right panels) Time series of the fractional area occluded by particles over the entire image (thick black line), and arriving in 10 subsampled areas in random, non-overlapping locations, for a “beam” diameter of (a) 0.7 cm, (b) 1.3 cm, and (c) 2.7 cm. The thin black line shows the mean of these 10 simulations, and gray shading indicates their standard deviation. (d–f) The corresponding root mean square deviation for each case.



sizes and morphologies (e.g., Durkin et al. 2015, 2021), which are likely to have second-order influences on  $\text{POC}:\text{atn}_c$  and  $\text{POC}:\text{atn}_{K_d}$ , similar to observations of  $\text{POC}:c_p$  in the water column (Cetinić et al. 2012). Decreased uncertainty in the predictions can be achieved on a site-specific basis if appropriate calibration data are collected (for instance, see differences among sites in Fig. 2). The relatively small size of the data set presented here did not allow rigorous testing of model parameterizations beyond the simple linear models used here, but as more comparison data spanning a broad range of oceanic properties are collected, this will become more feasible. Finally, the strong relationship between  $\text{atn}_c$  and  $\text{atn}_{K_d}$  observed here shows that measurements from different types of attenuation flux sensors can be quantitatively related to one another, facilitating meta-analyses across different studies.

### Beam transmissometers as $F_{\text{atn}_c}$ sensors

The main advantage to the use of a beam transmissometer as an OST is the ubiquity of the sensor, which has been commercially available for decades (e.g., C-Rover 2000 described previously; C-Star, Sea-Bird Scientific; LISST-Tau, Sequoia Scientific) and can be powered and sampled from commercially available, autonomous platforms. While not one of the standard sensors on profiling floats in the global Biogeochemical-Argo program (Bittig et al. 2019), beam transmissometers have been more widely deployed than any other  $F_{\text{atn}}$  sensor to date (Bishop et al. 2004; Estapa et al. 2013, 2019b; Terrats et al. 2020). Even though there are clear issues stemming from the fact that transmissometers were never designed to operate as sediment traps, they are well-characterized and relatively easy to acquire and deploy, adding great value to the  $F_{\text{atn}_c}$  data that these sensors continue to provide.

Because beam transmissometers are designed to measure the beam attenuation coefficient  $c$  in a water volume, they have characteristics that introduce bias when the sensor is instead employed as an OST. The primary issues stem from the small cross section of the light beam, and the upper pressure housing which blocks particles from sinking straight downward onto the detector window. The narrow beam cross section increases the time variability of the signal, and makes it more likely that accumulated particles will be dislodged in between samples leading to “negative jump” artifacts in the data. The algorithm that is used to process the raw  $\text{atn}_c$  vs. time data from a beam transmissometer-based OST must be tuned to eliminate these artifacts (“Sensitivity tests of the  $F_{\text{atn}_c}$  data processing algorithm” section). In addition, averaging periods of multiple days may be needed to compensate for variability caused by discrete particle arrivals into a small sampling cross-sectional area (Fig. 8), which limits the utility of a transmissometer-OST for looking at high-frequency variability in the GCP. Each combination of sensor, platform, and sampling design (i.e., the frequency with which the transmissometer is sampled) should be coupled to a round of sensitivity tests of the data processing algorithm, to ensure that the

thresholds and averaging periods are tuned to the raw data characteristics.

The upper pressure housing of a beam transmissometer blocks a fraction of sinking particles from reaching the detector surface. The main controls on this are the ambient turbulence which moves particles sideways into the sensor volume, and the sinking speeds of the particles. Dissipation rates in the upper kilometer of the global ocean range conservatively from  $10^{-11}$  to  $10^{-7} \text{ m}^2 \text{ s}^{-3}$ , with the lower end of this range characterizing deep waters far from any boundaries, and the middle part of the range ( $10^{-10}$ – $10^{-8} \text{ m}^2 \text{ s}^{-3}$ ) being more typical of the upper mesopelagic depths where GCP measurements are often made (Waterhouse et al. 2014; Whalen et al. 2015). Under the lowest- $\epsilon$  conditions, at most about 23% of particles are shaded, with a bias against collection of the fastest-sinking particles. Even at  $\epsilon > 10^{-8} \text{ m}^2 \text{ s}^{-3}$ , about 5% of particles will be shaded, although the sensitivity to particle speed diminishes (Fig. 7). Possible evidence for this in our observations is the departure from linearity of the relationship between  $F_{\text{atn}_c}$  and  $F_{\text{POC}}$ , relative to  $F_{\text{atn}_{K_d}}$  and  $F_{\text{POC}}$  (Fig. 2; Table 3). At higher  $F_{\text{POC}}$ , the relative decrease in the  $F_{\text{atn}_c}$  response compared to  $F_{\text{atn}_{K_d}}$  is consistent with increasing self-shading at higher fluxes, assuming that particles sink faster when  $F_{\text{POC}}$  is higher. Although dissipation rate measurements are not routine during BCP studies, global compilations of  $\epsilon$  estimates (e.g., Waterhouse et al. 2014; Whalen et al. 2015) should be consulted during study design to identify depths and regions where extra shading bias might impact  $F_{\text{atn}_c}$  measurements from a transmissometer-OST.

The empirical  $\text{atn}_c:\text{POC}$  calibration provided here takes into account the biases of transmissometer-based attenuation flux measurements and provides POC flux estimates with uncertainties of approximately  $19 \text{ mg-C m}^{-2} \text{ d}^{-1}$  ( $1.6 \text{ mmol-C m}^{-2} \text{ d}^{-1}$ ), for deployments under conditions that are represented by the studies compiled here. The data used in the calibration were collected at depths  $\leq 500 \text{ m}$ , all with C-Rover 2000 transmissometers. Applications of this calibration to  $F_{\text{atn}_c}$  observations collected at deeper depths away from boundaries may result in  $F_{\text{POC}}$  fluxes that are biased slightly low (due to more sensor self-shading under lower turbulence than is accounted for by the calibration). The uncertainty in the  $\text{atn}_c:\text{POC}$  flux calibration given here is large enough that transmissometer-derived flux estimates are best utilized in high-productivity, high-flux settings.

### Diffuse attenuation flux from transmitted-light, imaging sensors

Transmitted-light imaging sensors can provide diffuse attenuation flux measurements that also perform well as a proxy for POC flux. Although the  $\text{atn}_{K_d}:\text{POC}$  calibration has higher uncertainty than the one based on  $\text{atn}_c$ , imaging sensors provide complementary, simultaneous observations of particle size and identity, and they tend to have larger

collection areas than transmissometer-OSTs, which eliminates some of the temporal variability introduced by small beam cross sections. Carbon fluxes are often predicted from the analyzed images by assuming relationships among particle volume, carbon content, and composition (e.g., Durkin et al. 2021). In addition to these assumptions, the specific particle detection steps in the image analysis can introduce some uncertainty to the POC flux prediction (Giering et al. 2020b). POC flux prediction from  $F_{\text{atn}_{K_d}}$ , on the other hand, does not require particle detection or assumptions about particle properties, and thus can provide an important constraint for those more complex models.

In data sets where both were measured, we found that  $F_{\text{atn}_{K_d}}$  was consistently lower than  $F_{\text{atn}_c}$ , with an  $\text{atn}_{K_d}/\text{atn}_c$  ratio of about 0.4 (Fig. 2, bottom panel). This is analogous to the general relationship between the beam and diffuse attenuation coefficients in the water column, where  $K_d$  approaches  $c$  as the light field becomes less diffuse (Mobley 2022). In fact, Gordon et al. (1975) similarly predicted a  $K_d/c$  ratio of approximately 0.4 for typical oceanic particle scattering properties, under natural, diffuse light in the ocean. While the illumination and sensor light acceptance angles were not available for the variety of imaging systems represented in our  $\text{atn}_{K_d}$  data sets, the fact that we find a consistent  $\text{atn}_{K_d}/\text{atn}_c$  ratio here suggests that the cameras used to collect the data probably had fairly wide acceptance angles and illumination light sources were diffuse. This also implies that the  $\text{atn}_{K_d} : \text{POC}$  calibration reported here should not be applied to measurements from imaging sensors with very different geometries such as those using side-scattered light (e.g., Underwater Vision Profiler; Picheral et al. 2010, 2022) or collimated light (e.g., Laser Optical Plankton Counter or In Situ Ichthyoplankton Imaging System; Checkley et al. 2008; Cowen and Guigand 2008).

### Choosing the sampling area for $F_{\text{atn}}$ sensors

It makes intuitive sense that fluxes derived from a small sample area such as those on C-Rover transmissometers ( $\sim 0.7$  cm diam) are more sensitive to stochastic events such as the arrival (or loss) of large particles. By simulating expected attenuation timeseries from randomly selected subregions of a gel trap with a fairly large total imaged area ( $\sim 15$  cm<sup>2</sup>) we showed that this sensitivity improves markedly as the “beam” area is increased. Since this analysis was performed for particles larger than 300  $\mu\text{m}$ , we were not able to investigate results that include small particles. Assuming that smaller particles arrive steadily over time, they would have a buffering effect, reducing the overall sensitivity of the cumulative  $F_{\text{atn}_c}$  to episodic large particle arrivals. Thus, the image area required to adequately represent true particle variability is likely inversely proportional to the slope of the particle size distribution.

### Comments and recommendations

The choice of imaging vs. non-imaging  $F_{\text{atn}}$  sensors will depend on a number of application-specific considerations,

such as platform power and longevity, the need for ancillary particle size and identity information, and automated vs. manual data interpretation requirements. However, some universal considerations could help inform the design of future sensors for attenuation-based measurements of gravitational POC flux. Sensors that are unshaded from above will eliminate a key source of bias. The good relationship here between  $\text{atn}_c$  and  $\text{atn}_{K_d}$  suggests that a slightly oblique and/or diverging light source will still allow the precise determination of attenuation flux, and of POC flux with appropriate calibration data. Alternatively, self-shading can also be minimized by placing an attenuation sensor beneath a conical sediment trap where the light source occupies a minimal fraction of the collection area (such as the SES and CFE systems; “California Coastal (Sta. M)” and “California Coastal (CCE Long-Term Ecological Research)” sections; McGill et al. 2016; Bishop et al. 2016). However, such systems can be larger and more complex, perhaps not lending themselves to lower-cost, autonomous, mobile platforms. Finally, sampling areas of optimized OSTs should be large enough to minimize the uncertainty in attenuation flux measured at higher time resolution.

### Conclusion

There is a growing need for well-calibrated, autonomous measurements of POC flux in the ocean, which can support studies of the BCP’s mechanisms, evolution under climate change, and impact on Earth’s carbon cycle. Such measurements are also key components of any program that aims to enhance the ocean’s carbon dioxide uptake by accelerating the BCP. We have shown that the beam attenuation flux is a quasi-inherent optical property that can be used as a quantitative proxy for POC flux, much in the same way that the beam attenuation coefficient predicts the POC concentration. This proxy is based on a global calibration data set spanning three orders of magnitude in flux, from a broad range of ocean biogeochemical settings. We have also showed how diffuse attenuation flux, obtained here from images of accumulated particles, is related to beam attenuation flux and also predicts POC and mass flux. The most ubiquitous source of attenuation flux data at present comes from beam transmissometers used as OSTs, and here we have quantified their biases due to self-shading and a small beam cross section, to aid in the interpretation of their derived flux estimates. Finally, to aid in the broader use and development of optical POC flux sensors, we have outlined the key features that should be included in new sensor designs.

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